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To: Rulemaking File
Rulemaking Reference Number R-02-04
Mercury Waste Classification and Management

Subject: External Scientific Peer Review Report
California State University, Chico

Attached is the external scientific peer review report concerning the scientific bases for the Mercury Waste Classification and Management regulations that was performed by the California State University, Chico Research Foundation, in accordance to the requirements in Health and Safety Code section 57004, subdivision (d). The Mercury Report (August 2002), which contained the scientific bases, was reviewed by:

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Sections 1, 2, 3 and 5 of the Mercury Report, which was finalized in August 2002, contain the scientific bases for the regulations. Pursuant to Health and Safety Code section 57004, subdivision (d), the Department of Toxic Substances Control (DTSC) requested the reviewer to review and comment on Sections 1, 2, 3, and 5 of the Mercury Report (August 2002).

Dr. Wolfe's comments appear in **red** in the attached pages of the Mercury Report (August 2002). Dr. Wolfe did not submit all pages of the Mercury Report (August 2002) because DTSC only requested peer review of sections 1, 2, 3 and 5. Additionally, based on her expertise, Dr. Wolfe chose to only comment on sections 1, 2 and 5. However, DTSC requested another external peer review by Dr. A. Russell Flegal, University of California, Santa Cruz, who has commented on sections 1, 2, 3 and 5. Thus, all scientific bases of the regulations have been evaluated in accordance with Health and Safety Code section 57004, subdivision (d).

*The energy challenge facing California is real. Every Californian needs to take immediate action to reduce energy consumption.
For a list of simple ways you can reduce demand and cut your energy costs, see our Web-site at www.dtsc.ca.gov.*

MERCURY REPORT

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The energy challenge facing California is real. Every Californian needs to take immediate action to reduce energy consumption. For a list of simple ways you can reduce demand and cut your energy costs, see our Web-site at www.dtsc.ca.gov.

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During the development of this report, other individuals, who are not listed, were called upon during this project to provide information. DTSC would like to use this opportunity to acknowledge them as well.

Preface: Mercury Report August 2002

Introduction

The Mercury Report, August 2002, represents the finalized version of the Department of Toxic Substances Control's (DTSC's) earlier report, Draft Mercury Report (October 2001). The Mercury Report examines the problem of mercury contamination in California's environment and the contribution of the disposal of mercury-containing wastes not currently regulated as hazardous wastes. To fully consider the impacts of the hazardous waste identification and management options listed in Section 6 of the Draft Mercury Report, additional data was requested during the public workshops for the Proposed Regulation of Mercury-Containing Wastes, which were held between November 2001 and January 2002. Some additional data and information were received from the public workshops, but DTSC's conclusion that additional controls are necessary to protect public health and environment (Section 5) remains unchanged.

Summary of Comments and Revisions

The majority of comments that were received suggested various methods and strategies to reduce mercury emissions to the environment and affected Section 6, Options to Reducing the Amount of Anthropogenic Mercury Released to Land. Although Section 6 has been revised to reflect the regulatory concept to identify intentionally added mercury-containing products as a hazardous waste when they are discarded, the majority of these comments have not incorporated as revisions to Section 6. Instead, they have been considered in the proposed regulations for mercury that were public noticed on August 16, 2002. For further information on the proposed mercury regulations, please visit DTSC's website at www.dtsc.ca.gov.

Other comments received provide some recent data and information on the mercury trends and releases from anthropogenic sources. These were not incorporated into the August 2002 revisions. DTSC recognizes that the information provided reflects a national and global effort, both voluntary and mandatory, to decrease the use of mercury and to control mercury emissions from sources. However, DTSC's conclusion that additional controls are necessary to protect public health and environment (Section 5) by regulating mercury-containing wastes as hazardous wastes remains unchanged.

Air emission information in Section 3 was revised by the California Environmental Protection Agency's Air Resources Board (ARB). Data from ARB that affect text throughout the Draft Mercury Report, October 2001, have been similarly revised, specifically in Section 5. In addition, Section 3 was revised to include a technical correction, a reference to Assembly Bill 1760 (Chapter 849, Statutes of 1991) regarding removal of hazardous components from appliances.

Table of Contents

	page
Acknowledgments	I
Preface	ii
Table of Contents	iii
Tables and Figures	vii
Acronyms	ix
Executive Summary	1
Section 1:	
Nature and Extent of California’s Mercury Contamination: A Summary	7
I. Introduction	7
A. Properties of Mercury	7
B. Mercury Uses	7
C. Health Effects and Public Health	8
1. Health Effects	8
2. Public Health	8
D. Environmental Issues	9
1. Bioaccumulation and Biomagnification	9
2. Persistence	9
3. Mobility	9
II. Land Burden	10
A. Background Mercury Levels	10
B. Mercury-Containing Waste	10
1. Hazardous Waste Criteria	10
2. Disposal Options for Mercury-Containing Waste	11
C. Landfill Deposition of Mercury	12
1. Annual Disposal of Non-Hazardous Mercury-Containing Waste – Two Estimates	12
2. Leaching of Mercury from Landfills	14
D. Mercury Contaminated Sites in California	15
1. CalSites Data	15
2. Tailing Dumps	15
III. Mercury in California’s Air	16
A. Ambient Air Concentrations of Mercury in California	16
B. California Air Toxics Programs	16
1. The Toxics Air Contaminant Program (AB 1807)	16
2. The Air Toxics “Hot Spots” Program (AB 2588)	17
3. The Children’s Environmental Health Protection Program (SB 25)	17
C. Mercury health Data Associated with Air Exposures	18
D. Occupational Exposure Standards	18
E. Air Emissions	19
IV. Water Mercury Burden	20
A. Background/Ambient Water Quality	20
B. Standards	21

	page
1. Types of Water Quality Goals	21
2. Total Maximum Daily Loads (TMDLs) for Mercury	24
C. Water Mercury Sources	24
V. Public Health/Environmental Issues	24
A. Methylmercury in Fish/Consumption Advisories.....	25
B. Mercury Contaminated Sites	26
C. Nontraditional Sources of Mercury	26
Section I Key Points.....	27

Section 2:

Mercury's Chemistry and Toxicology – Human and Environmental Hazards	31
I. Introduction.....	31
II. Physical and Chemical Properties of Mercury and Mercury Compounds	31
A. Melting Point, Volatility	31
B. Covalent Bonding with Carbon.....	31
C. Important Mercury Compounds.....	32
D. Solubility of Mercury and Mercury Compounds.....	32
E. Unique Properties	32
III. The Global Mercury Cycle – Mercury Environmental Fate and Transport	32
A. The Global Mercury Cycle (Environmental Mercury Fluxes)	32
B. Fate and Transport of Mercury.....	33
1. Atmospheric.....	33
2. Terrestrial.....	34
3. Fresh Waters	34
4. Marine Waters	35
IV. Toxicology of Mercury and Mercury Compounds	37
A. Elemental Mercury	37
1. Toxicokinetics	37
2. Toxic Effects	37
3. Reference Exposure Standards.....	38
B. Mercuric Mercury	39
1. Toxicokinetics	39
2. Toxic Effects	40
3. Reference Exposure Standards.....	40
C. Methylmercury.....	41
1. Toxicokinetics	41
2. Toxic Effects	41
3. Reference Exposure Standards.....	42
4. Bioaccumulation	42
Section 2 Key Points.....	44

Section 3:

Sources of Mercury in California's Environment.....	49
I. Introduction.....	49
II. Natural Sources.....	49

	page
III. Anthropogenic Sources	49
A. Air Emission Sources In California	49
1. Paved and Unpaved Road Dust	50
2. Windblown Dust	50
3. Industrial Processes.....	50
4. Electrical Utilities.....	50
5. Petroleum and Related Products Manufacturing	51
6. Other Mobile Sources	51
7. Agricultural and Rangeland Prescribed Burning	51
8. Electric Lamp Breakage.....	51
9. On-Road Mobile Sources.....	51
10. Fuel Combustion Sources.....	51
11. Other Sources of Mercury Air Emissions	52
B. Temporal and Spatial Variability of Mercury Air Emissions	52
C. Water Mercury Sources	52
1. Past Activities – Legacy Waste	52
2. Current Activities.....	53
D. Land Mercury Sources	53
1. Past Activities	53
2. Current Activities.....	54
Section 3 Key Points.....	58

Section 4:

Mercury-Containing Products, Uses, and Alternatives	61
I. Introduction.....	61
II. Mercury-Containing Products and Alternatives.....	61
A. Measurement Devices—Temperature	61
1. Alternatives	61
B. Measurement Devices—Pressure.....	62
C. Electrical Devices – Switches and Thermostats.....	62
1. Alternatives	63
D. Dental, Medical, and Laboratory	63
1. Alternatives	63
E. Fungicides, Mildewicides, and Pesticides	64
F. Lighting.....	64
1. Alternatives	65
G. Household Batteries	65
1. Alternatives	66
III. Tables.....	66
Section 4 Key Points.....	68

	page
Section 5:	
Waste Contribution to the Mercury Environmental Burden.....	71
I. Introduction.....	71
II. Mercury Anthropogenic Sources and Emissions	71
A. Anthropogenic Sources – Raw Material	71
1. Domestic Supply Trends.....	71
2. Domestic Consumption (Demand) Trends.....	72
3. Mercury Flow Trends	75
B. Air Emissions	75
C. Water Emissions (Sources).....	77
D. Land Emissions (Disposal).....	77
E. Fluorescent Lamp Data.....	80
F. Dentistry.....	81
G. Data Limitations	81
III. Mercury Environmental Burden Assessment.....	82
A. Air and Water Waste Burden Assessment	82
B. Land Burden Assessment	84
Section 5 Key Points.....	86
Section 6:	
Options to Reducing the Amount of Anthropogenic Mercury Released to Land.....	91
I. Introduction.....	91
II. Background	91
III. Hazardous Waste Identification Options.....	92
A. Waste Types and Products	92
B. Hazardous Waste Identification Options	94
1. Regulate Intentionally Added Mercury-Containing Consumer Products When They Are Discarded as Hazardous Wastes	94
2. Regulate All Mercury-Containing Waste as Hazardous Waste	95
3. Regulate All Waste with Intentionally Added Mercury as Hazardous Waste	96
4. Develop a New Hazardous Waste Regulatory Threshold Number	96
5. Status Quo.....	97
IV. Hazardous Waste Management Options	98
A. Waste Types and Product Estimated Volumes and Capacities.....	98
B. Hazardous Waste Management Options	99
1. Universal Waste Management	99
2. Hazardous Waste Management	100
3. Phased Implementation	100
4. Landfill Disposal - Class I.....	101
5. Landfill Disposal - Class I, II, or III	101
V. Options Limitations	102
VI. Recommendation	102
Section 6 Key Points.....	103

	page
Appendix A:	
Summary of Nationwide Mercury Efforts	106
References	114

Figures and Tables

	page
Section 1:	
Nature and Extent of California’s Mercury Contamination: A Summary	
Table 1-1: OEHHA/ARB Approved Risk Assessment Health Values	18
Table 1-2: Industrial Hygiene Limits for Occupational Exposure (mg/m ³)	
Mercury Inhalation	19
Table 1-3: Summary of Water Quality Goals in California.....	21
Table 1-4: Summary of State and Federal Water Quality Standards for Mercury	23
Table 1-5: Sport Fish Consumption Advisories for Mercury Contaminated	
Water Bodies, 1999	25
Section 2:	
Mercury’s Chemistry and Toxicology – Human and Environmental Hazards	
Table 2-1: Physical and Chemical Properties of Selected Mercury Species	31
Table 2-2: Estimated Mercury Content of Environmental Media – Worldwide	32
Table 2-3: Environmental Influences on the Rate of Methylation of	
Aquatic Mercury.....	35
Table 2-4: Significant Methylmercury Inputs to the World’s Coastal Waters.....	36
Table 2-5: Reference Doses (RfDs) and Reference Concentrations (RfCs)	
for Mercury, Elemental	39
Table 2-6: Minimal Risk Level (MRLs) for Mercury, Metallic – March 1996	39
Table 2-7: Reference Doses (RfDs) and Reference Concentrations (RfCs)	
for Mercuric Chloride	40
Table 2-8 Minimal Risk Level (MRLs) for Mercury, Inorganic – March 1996.....	40
Table 2-9: Reference Doses (RfDs) and Reference Concentrations (RfCs)	
for Methylmercury	42
Table 2-10 Minimal Risk Level (MRLs) for Methylmercuric Chloride – March 1996.	42
Section 3:	
Sources of Mercury in California’s Environment	
Table 3-1: Air Emissions of Mercury in California in 2000.....	50
Section 4:	
Mercury-Containing Products, Uses, and Alternatives	
Table 4-1: Some Mercury Compounds and Uses	66
Table 4-2: Mercury Uses in Products.....	66
Table 4-3: Major Mercury-Containing Products and Alternatives	67

Section 5:

Waste Contribution to the Mercury Environmental Burden

Figure 5-1: U.S. Industrial Reported Consumption of Mercury (1970-1997)	73
Figure 5-2: Apparent Supply and Reported Consumption of Mercury (1970-1998) .	74
Table 5-1: U.S. Mercury Emissions from Combustion Sources, 1996	75
Table 5-2: California Waste Derived Air Emissions for 2000	76
Table 5-3: Discards of Mercury in Products in the Municipal Solid Waste Stream 1970 to 2000 (in short tons)	78
Table 5-4: Discards of Mercury in Products in the Municipal Solid Waste Stream 1970 to 2000 (In Percent of Total Discards)	79

Section 6:

Options to Reducing the Amount of Anthropogenic Mercury Released to Land

Table 6-1: Waste Types / Products	92
Table 6-2: Waste Types / Products – Estimated Volumes and Capacities	98

Definition of Acronyms and Abbreviations

μg/m ³	microgram per cubic meter
μg/L	micrograms per liter
22 CCR	Title 22, California Code of Regulations
AB	Assembly Bill
AB 1807	Toxics Air Contaminant Program
AB 2588	Air Toxic "Hot Spots" Program
APCD	Air Pollution Control District
AQMD	Air Quality Management District
ARB	California Air Resources Board
ATSDR	Agency for Toxic Substances and Disease Registry
CEIDARS	California Emission Inventory Development and Reporting System
CERCLA	Comprehensive Environmental Response, Compensation, and Liability Act
CH ₃ Hs	Methyl mercury
CIWMB	California Integrated Waste Management Board
CTR	California Toxics Rule
DHS	Department of Health Services
DTSC	Department of Toxic Substances Control
Hg ⁰	Elemental mercury, metallic mercury, quicksilver
HgS	Inorganic (oxidized) mercury
HID	High Intensity Discharge
HPSuL	High-Pressure Sulfur Lamps
HSC	Health and Safety Code
IRIS	Integrated Risk Information System
LOAEL	Lowest Observed Adverse Effects Level
LPSL	Low-Pressure Sodium Lamps
MCL	Maximum Contaminant Level
MCLG	Maximum Contaminant Level Goal
mg/kg/day	milligram per kilogram per day
mg/L	milligrams per liter
MRL	Minimal Risk Level
MSW	Municipal Solid Waste
NAS	National Academy of Sciences
ng/g	nanograms per gram
ng/L	nanogram per liter
ng/m ³	nanogram per cubic meter
NOAEL	No Observed Adverse Effects Level
NPL	National Priorities List
OEHHA	California's Office of Environmental Health Hazard Assessment
OSHA	Occupational Safety and Health Administration
PEL	Permissible Exposure Limit

PHG	Public Health Goal
POTW	Publicly Owned Treatment Works
RCRA	Resource Conservation and Recovery Act
RELS	Reference Exposure Levels
RfC	Reference Concentration
RfD	Reference Dose
RWQCB	Regional Water Quality Control Boards
RWQCP	Regional Water Quality Control Plant
San Francisco Bay TMDL 2000 Report	Abu-Saba, et al., <i>Watershed Management of Mercury in the San Francisco Bay Estuary: Total Maximum Daily Load Report to U.S. EPA</i> , June 2000
SARA	Superfund Amendments and Reauthorization Act
SB	Senate Bill
SB 25	Children's Environmental Health Protection Program
SIC	Standard Industrial Classifications
SNARLs	Suggested No Adverse Response Levels
SWAT	Solid Waste Assessment Test
SWRCB	State Water Resources Control Board
TAC	Toxic Air Contaminant
TCLP	Toxicity Characteristic Leaching Procedure
TMDL	Total Maximum Daily Load
TTLC	Total Threshold Limit Concentration
U.S. EPA	United States Environmental Protection Agency
USEPA 1992 Study	U.S. EPA 1992. <i>Characterization of Products Containing Mercury in Municipal Solid Waste in the United States, 1970 to 2000</i>
USEPA 1997 Study	U.S. EPA 1997. <i>Mercury Study Report to Congress</i>
USGS	United States Geological Survey
USGS 2000 Study	United States Geologic Survey, June 2000
WET	Waste Extraction Test
WMUD	Waste Management Unit Database

Executive Summary

This report examines the problem of mercury contamination in California's environment and the contribution of the disposal of mercury-containing waste not currently regulated as hazardous waste. The report consists of six sections. The first provides a general overview of California's mercury problem, while each of Sections 2 through 5 focuses on a different aspect of mercury in more detail. The final section examines several options for reducing the further contamination of California's land, and recommends changes to the State's criteria used to classify mercury-containing waste as hazardous waste.

Section 1 provides a general overview of mercury in the State's environment. Mercury is a metal that occurs naturally in California; its use has been and continues to be widespread throughout the world. As a result, mercury contamination is found throughout the State, in all environmental media. This widespread contamination is especially serious because of mercury's unique combination of properties.

Because metallic mercury is a liquid at room temperature, it is especially mobile in the environment. It is also persistent in the environment, and forms organomercuric compounds that can bioaccumulate in organisms and biomagnify in the food web. High-level predators can have mercury body burdens that are several orders of magnitude higher than the concentrations found in the surrounding environment. Environmental mercury can readily move among environmental media. For example, mercury that is emitted directly to air is inevitably deposited on land and water. Similarly, mercury contained in waste that is deposited in municipal landfills can dissolve in landfill leachate and potentially contaminate the State's waters.

Mercury's health and environmental hazards have led to the development of numerous regulatory standards for mercury in waste, air, and water, as well as occupational exposure standards. These standards have been exceeded in some cases, necessitating action by responsible parties, as well as State and federal agencies. A number of sites in California are sufficiently contaminated with mercury to make clean-up or other mitigation activities necessary. Similarly, some of the State's water bodies exceed water quality standards for mercury, triggering a requirement under the Federal Clean Water Act that Total Maximum Daily Loads (TMDLs) be developed. Sport fish in certain of State's water bodies are sufficiently contaminated with methylmercury that the Office of Environmental Health Hazard Assessment (OEHHA) has advised the public to restrict or eliminate consumption of them.

Under current hazardous waste identification criteria, some mercury-containing waste is sometimes classified as nonhazardous waste, and consequently, it legally may be disposed in municipal landfills. While the mercury concentration in such waste is relatively low, the large volume of waste that is disposed contributes a significant amount of mercury to municipal landfills. Studies have shown that municipal landfills can leak detectable concentrations of mercury and, in a recent study, various mercury

species were found in municipal landfill gas.

Section 2 describes mercury's chemistry and toxicology. Three important forms of mercury exist in the environment: metallic mercury, mercuric mercury, and methylmercury. Each has distinct chemical and physical properties, environmental behavior, and toxicology. Mercury's environmental fate and transport are described in terms of flux or movement between environmental media. Up to 75 percent of the mercury emitted to the world's atmosphere is of anthropogenic origin, and the world's atmospheric mercury load has increased between two and five-fold since industrialization.

Atmospheric mercury is ultimately deposited on land or water, either in precipitation or via dry deposition of particulates. Of the environmental media, mercury is least mobile in soil. However, mercury can form soluble complexes with organic ligands in soil, and subsequently dissolve in runoff or leach from municipal landfills. Mercury that enters marine environments can be methylated by both biotic and abiotic processes. It can enter the marine food web via plankton in the water column and via larger invertebrates in marine sediments.

Eighty percent of inhaled elemental mercury is absorbed into the body. Neurotoxic effects are the most sensitive toxicological endpoint of elemental mercury. They include tremors, changeable emotional state, insomnia, headaches, sensory loss, memory loss and impaired cognitive function.

Mercuric mercury enters the body via inhalation, ingestion, or dermal exposure, and can be methylated by gastrointestinal microbes. Renal toxicity is the most sensitive toxic endpoint in humans.

Methylmercury can be absorbed by the lungs and is well absorbed in the digestive tract. Humans absorb 95 percent of the methylmercury in the fish they consume. Methylmercury is lipophilic and readily crosses the blood brain and placental barriers. Methylmercury's half-life in blood is estimated to be 50 days and is a potent developmental and neurological toxin in humans.

Inorganic and elemental mercury are both toxic, but of the environmentally important forms, methylmercury poses the greatest risk to human health and the environment due to its high toxicity and the fact that it bioaccumulates in aquatic organisms. Consumption of contaminated fish is the primary route of human methylmercury exposure in humans.

Section 3 discusses the sources of mercury in California's environment. The mercury in the State's environment originates from both natural and human sources. Both historical and ongoing sources have added to California's current environmental mercury burden. Important historical mercury sources include gold and mercury mining and past waste and industrial management practices, such as open garbage burning; and the collection of industrial process wastes in unlined sumps, ponds, and lagoons.

Mercury released into the environment from these and other human activities continues to move in the global mercury cycle.

California's mercury air emissions totaled approximately 20 short tons in 2000. Some of the notable sources were windblown dust, geothermal energy production, cement manufacturing, petroleum-related manufacturing, electric utilities, waste burning, and fluorescent tube breakage.

Publicly owned treatment works (POTWs) are current sources of small, but quantifiable mercury discharges to the State's waters. By far, the largest contributor of mercury to the State's waters is the legacy waste from past mining activities. Thousands of tons of mercury were lost to the State's environment from past placer gold mining. Drainage from more than 300 abandoned mercury mines and prospects found along the California Coast Range continues to release mercury to the region's waters.

Land disposal of mercury-containing wastes contributes to California's environmental mercury loading through direct land contamination, surface runoff, leaching to water, and, potentially, atmospheric emissions in landfill gas. A recent study of a Florida municipal landfill showed detectable amounts of mercury compounds in landfill gas, suggesting that landfill gas may be a larger source of mercury air emissions than was previously believed.

Mercury-containing wastes currently disposed in municipal landfills include fluorescent lamps, soils, industrial wastes, ashes, POTW sludges, and non-metallic components from shredded automobiles that are contaminated with mercury.

Section 4, discusses various mercury-containing products, their uses, and some mercury-free alternatives to these products. Mercury's physical properties, including its high density and liquid state at room temperature make it useful in mechanical switching devices, such as thermostats. Mercury is also used in thermometers, a variety of measurement devices, electrical devices, dentistry, medicine, lighting, and biocides. Despite the decrease in mercury consumption in most applications, releases to the environment are expected to continue as spent mercury-containing products are disposed. A growing list of viable alternatives to mercury-containing products is becoming available for most consumer applications.

Section 5 discusses the contribution of the disposal of waste to environmental mercury loading. Human activities have caused an estimated three-fold increase in the global environment mercury burden. However, in recent years, the use of mercury has been significantly curtailed. U.S. mine production and imports of mercury decreased rapidly from 1986 to 1992; by 1993, most of mercury in the market originated from secondary (recycled) sources. Domestic mercury consumption dropped from more than 2426 short tons in 1976 to less than 441 short tons in 1998.

A number of waste management activities, including waste combustion, are sources of mercury emissions to air. In 1994 and 1995, approximately 87 percent of the nation's

atmospheric mercury emissions originated from combustion point sources. These sources included fossil fuel combustion, which emitted 84 short tons of mercury to the nation's air in 1996, and waste combustion and incineration, which contributed 60 short tons. California's mercury air emissions from waste management activities, including combustion and landfill sources, were 2.24 tons in 2000, with 370 pounds attributed to broken fluorescent tubes.

A large proportion of California's aquatic mercury load originates from legacy waste from inoperative mercury and gold mines. Other waste sources include leaching and runoff from landfills, atmospheric deposition, and the sewer system. It is estimated that 1180 pounds of mercury from dental offices is present in water entering the State's POTWs for treatment. POTWs typically remove 90 percent of the mercury from their influents. At this rate, 118 pounds of the dental mercury would be discharged to California's waters. The San Francisco Bay Regional Water Quality Control Board has estimated that, annually, between 22 and 286 pounds of mercury from fluorescent lights potentially enters the San Francisco Bay alone.

The USGS estimated that the amount of mercury disposed in landfills fell from 832 short tons in 1990 to 325 short tons in 1996. Mercury from household batteries and lighting comprise of the majority of the discards in the municipal solid waste stream from 1970 to 1989 and was projected to be the same in 2000. U.S. EPA's study showed that the mercury contribution from fever thermometers and thermostats did not show signs of decreasing between 1970 and 1989, and no significant reductions were projected for 2000.

The mercury content of fluorescent lamps decreased sharply between 1985 and 1995, but the rate of reduction has decreased in recent years. Without affecting their life, further reductions in the mercury content of lamps may be increasingly difficult for the industry to achieve. U.S. EPA estimates that 26.7 tons of mercury was disposed in electric lights, nationally, in 1989, while California estimates that 1.3 short tons of mercury from fluorescent lamps will be disposed in 2001. California dentists generated an estimated 2.2 tons of mercury from dental amalgam that was disposed or recycled in 2000. Automobiles potentially contribute 0.75 to 1.5 short tons of mercury to nonhazardous waste landfills per year through auto shredder waste. DTSC's Auto Shredder Initiative sampling and laboratory analyses showed that in 2001, approximately 0.93 tons of mercury was found auto shredder waste (resulting from shredding automobiles and appliances), and that 0.4 short tons originated from automobiles.

Anthropogenic mercury air emissions are decreasing as a result of decreases in industrial uses of the metal, as well as improvements in air pollution control devices. While the use of mercury has continued to drop, the environmental mercury load remains unacceptably high. This is evidenced by numerous sport fish consumption advisories, by the existence of mercury-contaminated sites, and by the numerous legislative and regulatory efforts to reduce mercury contamination.

The Department of Toxic Substances Control (DTSC) may recommend regulation of all mercury-containing waste as hazardous waste, in order to promote pollution prevention and recycling and to limit further environmental mercury loading.

Several options for reducing the amount of mercury released to the environment are outlined in Section 6. The promotion of pollution prevention, the use of mercury alternatives, and mercury recycling may be best accomplished by redefining the hazardous waste identification criteria for mercury. DTSC is recommending the regulatory concept to identify intentionally added mercury-containing products as a hazardous waste when they are discarded. Where appropriate, certain mercury-containing products could be managed under DTSC's universal waste management standards. Disposal of regulated mercury-containing products would be limited to Class I landfills. In order to facilitate compliance, development and identification of substitutes for mercury-containing products, and development of infrastructure, the implementation of the new mercury criteria would be phased in over time.

Other hazardous waste identification options that may be considered are variations of "listing" mercury-containing wastes and are as follows:

- Regulate all mercury-containing waste as hazardous waste
- Regulate all waste with intentionally added mercury as hazardous waste
- Develop a new hazardous waste regulatory threshold number
- Status quo

Hazardous waste management options may also be considered include the following:

- Universal waste management
- Full hazardous waste management standards
- Phased implementation
- Landfill disposal - Class I landfill
- Landfill disposal – Class I, II, or III

Additional data is needed in order to fully consider the impacts of the hazardous waste identification and management options listed above. Information that was received during the public workshops, which were held between November 2001 and January 2002, reflected various methods and strategies to reduce mercury released to land.

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Section 1: Nature and Extent of California's Mercury Contamination: A Summary

I. Introduction

Mercury is a toxic heavy metal that has been used for millennia because of its unique combination of chemical and physical properties. Mercury's widespread use and subsequent release into the environment, combined with its high toxicity, persistence in the environment, and propensity to bioaccumulate and biomagnify in the aquatic food web, make it a contaminant of special concern. Although the use of mercury has been curtailed nationwide, and regulatory standards have been established to limit its release to the environment, mercury continues to cause public health and environmental concerns. These are evidenced by fish advisories issued by California's Office of Environmental Health Hazard Assessment (OEHHA) for a number of California recreational waters.

This section provides an overview of the properties and uses of mercury, the environmental behavior and toxicity of different forms of the metal, and the origin and extent of the State's land, air, and water contamination. The report discusses the disposal of mercury-containing waste not currently regulated under the State's hazardous waste laws. It then focuses on State and federal regulatory standards for mercury in the various environmental media and in the workplace and instances when these standards have been exceeded.

A. Properties of Mercury

Elemental mercury is a liquid over a wide range of temperatures. It exists in a variety of chemical forms in the environment, each of which has distinct chemical and physical properties and toxicology. As it moves through different environmental media, mercury's chemical oxidation state can change. "Through natural chemical and biological reactions, mercury changes form among these species, becoming alternately more or less soluble in water, more or less toxic, and more or less biologically available."¹ Important forms of mercury in the environment include:

- Elemental or metallic mercury, also known as quicksilver (Hg^0),
- Inorganic (oxidized) mercury, including the ore cinnabar (HgS), and
- Organic mercury, including methyl mercury (CH_3Hg).

B. Mercury Uses

Elemental mercury is a liquid at room temperature, expands at a uniform rate with increasing temperature, is relatively dense, and has a low surface tension. These properties have made it very useful in measurement devices such as thermometers, manometers, and barometers. Because it conducts electricity, mercury is also used in a variety of electrical applications, such as electrical lights and switches. Mercury easily forms alloys, called amalgams, with many metals. This property has been exploited in several industries, notably dentistry, gold mining, and chemical manufacturing. Mercury has also been used as a fungicide, mildewicide and pesticide.

C. Health Effects and Public Health

1. Health Effects

Mercury is toxic in all its forms, but its routes of entry, mode of action, and potency are different for each of them. Mercury's toxicology is discussed in detail in Section 2 of this report, but the salient points are briefly summarized here.

Metallic mercury is poorly absorbed in the digestive tract, but readily enters the body via inhalation.² The toxic effects of metallic mercury on the central nervous system were known by the 19th century in occupational exposures. Mercury was extensively used in the production of felt, and persons who worked with felt were noted to behave strangely. The Mad Hatter in Lewis Carroll's 1865 novel Alice's Adventures in Wonderland exhibited symptoms of acute metallic mercury poisoning: excitability, delirium, and hallucinations.³ Metallic mercury toxicity is also characterized by tremors, blurred vision, speech problems, and excessive shyness. Mercury is also toxic to the gastrointestinal tract and the respiratory system.⁴

Inorganic mercury salts are relatively well absorbed in the digestive tract. After ingestion, inorganic mercury is distributed throughout the body in the bloodstream, but it concentrates in the kidneys.⁵ Inorganic mercury is toxic to the kidneys. In laboratory animal studies, ingestion of inorganic mercury led to increases in kidney weight and necrosis (death) of the proximal tubules.⁶

Organomercurics, of which methylmercury and dimethylmercury are two, are the most toxic mercury compounds. A Dartmouth University researcher died in 1997 after dermal exposure to a drop of dimethylmercury that passed through her glove.⁷

Methylmercury's extreme toxicity has been well documented in a number of epidemiological studies.

2. Public Health

The most infamous outbreak of mercury poisoning was first identified in 1956, among residents of the Minamata Bay region on the island of Kyushu, Japan. These people were highly exposed to methylmercury from ongoing, heavy consumption of fish, which were contaminated with mercury from industrial pollution. According to one author, 59 percent of 628 exposed persons exhibited mental or neurological disorders.⁸ Symptoms included tingling in the fingers and toes, difficulty grasping, walking, running, swallowing, and speaking and impaired vision and hearing. "Examination of the brains of severely affected patients that died revealed marked atrophy of the brain (55% normal volume and weight) . . ."⁹ Children born to exposed mothers had a high rate of birth defects, which included mental impairment, delayed development, and severe brain damage.

D. Environmental Issues

1. Bioaccumulation and Biomagnification

Metallic mercury (Hg^0) is converted to the extremely toxic and readily absorbed compound methylmercury by sulfur-reducing bacteria in the lower sediment layers of lakes, rivers, and streams. Unlike metallic mercury, methylmercury is readily absorbed and retained by organisms. This property results in an increase, over time, in the concentration of the methylmercury in aquatic organisms that live in contaminated waters—a phenomenon known as bioaccumulation. Fish take up methylmercury directly, across their gills¹⁰, and predatory fish and birds absorb much of the methylmercury that their prey have absorbed. Consequently, the predators at the highest levels of the food web have the highest concentrations of methylmercury in their bodies. Contaminants that become more concentrated as they move from organisms at lower trophic levels of the aquatic food web (prey) to organisms at higher levels (predators) are said to undergo biomagnification. **Piscivorous wildlife, unlike humans, are not protected by fish advisories, as they do not have alternate sources of food.**

2. Persistence

Heavy metals like mercury are believed to originate in supernovae¹¹, and can neither be created nor destroyed. The mercury that has been used by humans over more than two millennia was extracted mainly from deposits of cinnabar, the most common mercury ore. Mercury is also naturally present in coal, and is released to the environment when coal is burned.

Some of the mercury present in mineral deposits is gradually mobilized to air and water, but human activities to extract and use these resources have significantly increased the amount of mercury that is mobile in the environment.¹² One study estimates that since the beginning of the industrialized period, mercury emissions resulting from human activities have led to threefold increases in worldwide atmospheric and oceanic mercury concentrations.¹³ Once mercury is mobilized in the biosphere, it remains there and increases the exposures to humans and the environment.

3. Mobility

Due to its chemical and physical properties, mercury is mobile in the biosphere, both within and between environmental media (land, water, and air).¹⁴ The movement of mercury in the environment is greatly affected by its oxidation state and is described in terms of a global cycle, which will be discussed later in some detail in Section 3. Briefly, the mercury cycle describes the movement of mercury between land, air and water. Mercury is emitted directly to air by both natural and human activities. Some fraction of the airborne mercury is deposited to land or water near the source of emission, while the rest enters the global atmospheric cycle, and is transported worldwide.¹⁵ Once in the atmosphere, mercury can be deposited far from the emission source by two mechanisms: dry deposition and wet deposition (deposition in rain or snow).¹⁶ Atmospheric deposition can be either to land or to water. Mercury is also released directly to water and land by natural and human activities, and can migrate from water to air, and from land to air and/or water.¹⁷

II. Land Burden

Environmental mercury moves between soil, water, and air and originates from both natural and anthropogenic sources. While the metallic and inorganic forms of mercury most commonly found in soils are toxic in their own right, the especially toxic form methylmercury is found mainly in aquatic environments. The metallic and inorganic forms found in soils contribute to aquatic methylmercury loading, because they can migrate into surface waters, where they are readily converted to methylmercury by aquatic bacteria. Soil mercury levels have increased as a result of human activities. Consequently, the amount of mercury that is mobile in the environment has also risen, as have the risks to public health and the environment.

In order to prevent further increases in soil mercury loading, standards restricting the land disposal of mercury-containing waste have been established in regulations. Some of these standards are in the form of thresholds. Regulatory thresholds are calculated based on predefined levels of acceptable risk, using theoretical models of the behavior of the contaminant of concern. The models consider a contaminant's concentration, mobility, and toxicity, among other factors. Whether a mercury-containing waste exceeds the established thresholds for mercury determines where it may be disposed, and how it must be managed prior to disposal.

A. Background Mercury Levels

The earth's crust naturally contains small amounts of mercury. In some areas, soil mercury concentrations are elevated above typical background levels. The sources of such elevations vary, and are both natural and anthropogenic. Normally, soil parent materials' mercury content is quite low, and the soil that is formed from them is generally naturally low in mercury. In areas where mercury-rich minerals are abundant, higher soil mercury concentrations are observed. Andersson analyzed data for the mercury content of common soil-forming minerals from numerous published studies and found that igneous rock, coarser-grained soil fractions, sandstone, and limestone all typically have mercury concentrations below 50 nanograms per gram.¹⁸ Andersson states that "(a) normal range of 10-50 ng/g seems to be reasonable for soil parent material, but much higher levels may be found in certain areas."¹⁹ In its *Mercury Study Report to Congress* (U.S. EPA 1997 Study), the United States Environmental Protection Agency cites an estimate that typically, United States soils contain between 8 and 117 ng/g (dry weight) of mercury.²⁰

B. Mercury-Containing Waste

1. Hazardous Waste Criteria

Given that there is a range of background mercury levels in soil, regulations were adopted in the mid-1980s to control the disposal of mercury-containing industrial and consumer wastes in landfills. Both State and federal regulations contain criteria to determine whether a waste is hazardous, in order to determine its proper management and disposal. These criteria include threshold concentrations for leachable mercury; wastes that exceed the thresholds are considered hazardous and must be managed

accordingly. Both the federal Toxicity Characteristic Leaching Procedure (TCLP) and California's Waste Extraction Test (WET) are based on the principle that toxic substances such as mercury can dissolve in landfill leachate. Once dissolved, they can migrate from a disposal area and pollute ground or surface waters. While both procedures are designed to simulate the leaching of chemicals that are buried in a landfill, they differ in some respects, and the WET is generally considered more aggressive for inorganic chemicals. In both federal and State leaching procedures, mercury-containing waste is classified as hazardous when it has an extractable mercury concentration at or above 0.2 mg/L.

In California, wastes whose leachable mercury concentrations do not exceed the 0.2 mg/L threshold are nevertheless classified as hazardous if their total mercury concentration equals or exceeds 20 mg/kg. Such "Total Threshold Limit Concentrations" (TTLIC) have no counterparts in the federal waste classification scheme.

Whether or not their mercury concentrations exceed State and federal thresholds, certain "listed" wastes are classified as hazardous. U.S. EPA has established four hazardous waste lists in its regulations. Several listed wastes are included because they contain mercury.

The hazardous waste identification criteria determine what handling and disposal requirements apply to a waste. Mercury-containing waste that meets any of the criteria must be stored, transported, and disposed in a manner that is protective of public health and environment, in accordance with hazardous waste management standards found in federal and State regulations.

2. Disposal Options for Mercury-Containing Waste

Disposal options are limited for mercury-containing waste that meets hazardous waste identification criteria. In California, hazardous waste may only be disposed to land in Class I landfills, which are hazardous waste landfills. Class I landfills must meet stringent requirements to prevent migration of chemicals into the environment. They must be constructed with a protective liner, leachate collection system, and are subject to site-specific permitting requirements and waste acceptance criteria. The management of a Class I landfill is overseen by two state agencies: the Regional Water Quality Control Board (RWQCB) in whose jurisdiction it is located, and the Department of Toxic Substances Control (DTSC).*

A second category of landfills, Class II landfills, is designed to accept designated wastes—wastes whose land disposal may threaten water quality. Some Class II landfills accept municipal solid waste, but others are restricted from doing so. Typically, Class II landfills accept only lower risk hazardous waste. Before it may accept any hazardous waste, a Class II landfill must obtain a variance from DTSC, and must obtain

* Local Air Pollution Control Districts (APCDs) and Air Quality Management Districts (AQMDs) oversee air quality issues at landfills.

a permit from the local RWQCB and the California Integrated Waste Management Board (CIWMB). *

Class III landfills, also referred to as municipal solid waste landfills, are also overseen by RWQCB and CIWMB.* Because the design, siting, and permitting requirements for Class III landfills are less stringent than those for Class I and II landfills, hazardous substances are more likely to leach into the surrounding environment from Class III landfills. For this reason, they may generally accept only non-hazardous waste for disposal. In special instances, upon approval of DTSC, the local RWQCB, and CIWMB, Class III landfills may accept lower risk hazardous wastes, such as asbestos, treated wood and wastes containing solid metal.

If landfill gas controls are required by the air pollution control agencies, Class II and III landfills are designed to include landfill gas collection systems to either allow the landfill gas collected to be burned for energy recovery or flared. Class I landfills do not have landfill gas collection systems as they do not accept putrescible or volatile organic waste, which creates an environment to produce landfill gas.

C. Landfill Deposition of Mercury

1. Annual Disposal of Non-Hazardous Mercury-Containing Waste—Two Estimates
Mercury-containing waste that meets hazardous waste identification criteria is subject to more stringent management and disposal standards than is mercury-containing waste that does not meet the criteria. Disposal of nonhazardous waste that contains mercury in Class III landfills is a concern, due both to the less protective management standards for the waste prior to disposal, and the less stringent design and operation standards for the landfills.

a. United States Geological Survey (USGS) Estimate

In its study, *The Materials Flow of Mercury in the Economies of the United States and the World* (USGS 2000 Study), the USGS estimates the total amount of mercury deposited in U.S. municipal landfills. The estimates are based on data from 1994 and 1995, published in the U.S. EPA 1997 Study.²¹ USGS calculates the following values for nationwide landfill disposal of mercury:

- The total mass of mercury in municipal solid waste in the United States was 340 tons†.
- 299 tons of mercury were contained in waste that was directly disposed in municipal landfills.
- The remaining 41 tons of municipal waste were incinerated in municipal waste combustors.
- The average mercury-removal efficiency of various emission control devices²² used for municipal waste incinerators was determined to be 27 percent. This value was

† All references to 'tons' denote short tons. For consistency, all weight measurements were converted to short tons in this report. A short ton is 2000 pounds, or 0.907 metric tons. A metric ton is 1000 kilograms, or 2200 lbs.

used to calculate that approximately 11 tons of mercury were captured by these devices and subsequently deposited in landfills, while the remaining 30 tons were emitted to the atmosphere.

- The total amount of mercury going to landfills was calculated to be 310 tons. (340 tons – 41 tons + 11 tons = 310 tons.)

According to United States Census data, California's population represents approximately 12 percent of the total United States population.^{‡23} Assuming the per capita generation of municipal solid waste is approximately the same in California as in the United States, and using USGS's estimate of 310 tons of mercury disposed in landfills nationally, approximately 37.2 tons of mercury were disposed in ~~the~~ California's landfills in each of 1994 and 1995.

b. U.S. EPA Estimate

In the U.S. EPA 1997 Study, U.S. EPA estimates that 227.6 tons of mercury were discarded in the United States in 1995, and that 144.6 tons would be discarded in 2000.

This data was taken from a 1992 U.S. EPA study *Characterization of Products Containing Mercury in Municipal Solid Waste in the United States, 1970 to 2000* (U.S. EPA 1992 Study). The 1992 data was modified to account for federal legislation adopted in 1996 that restricted the use of mercury in batteries, which led to the elimination of mercury from most batteries.²⁴ The contribution of mercury in discarded fluorescent tubes to the total was also adjusted downward in the 1997 study to account for the reduction in the average amount of mercury used in their manufacture.²⁵

However, the 1997 study did not anticipate U.S. EPA's 1999 rulemaking that added mercury-containing hazardous waste lamps to the universal waste program, nor the inclusion of these lamps in universal waste regulations promulgated in many states, including California. Consequently, the U.S. EPA 1997 Study may overestimate the amount of mercury disposed into municipal solid waste landfills.

The amount of mercury disposed by Californians can be calculated from U.S. EPA's national data, and can be compared with the value calculated from USGS's estimate. Assuming, as before, that Californians accounted for 12 percent of total U.S. disposal, approximately 27.3 tons of mercury were disposed in the State's municipal landfills in 1995, and 17.3 tons would have been disposed in 2000.²⁶

The estimate based on the U.S. EPA 1997 Study is somewhat lower than the estimate based on the USGS 2000 Study. This difference may be explained by the fact that U.S. EPA's national estimate is based on the disposal of a list of mercury-containing products, while USGS's total is based on estimates of the loss of mercury from municipal waste combustors, some of which may originate from wastes not included in U.S. EPA's list. Nevertheless, the two values are in rough agreement.

[‡] According to the Census Bureau, in 1990, California had 29.76 million of the 248.7 million people in the United States. In 2000, the State's population was 33.87 million of the 281.4 million people in the nation. California's percentage of the nation's population has remained constant, at approximately 12 percent.

California adopted its Universal Waste Rule in 2000, which provided alternative management requirements for mercury-containing lamps and thermostats in order to encourage their proper management and diversion from non-hazardous waste landfills. Any decrease in the disposal of these items that may have resulted from these recently adopted regulations is not reflected in either estimate.

2. Leaching of Mercury from Landfills

Groundwater at municipal solid waste (Class III) landfills is currently monitored for mercury under the waste discharge requirements issued by the RWQCBs. If the concentration of any constituent of concern exceeds the corresponding Maximum Contaminant Level (MCL), enforcement is brought by the local RWQCB. The MCL for mercury is MCL 0.002 mg/L.

When U.S. EPA published its proposed rule on spent mercury lamps in 1994, the agency requested data on the mercury content of landfill leachates or groundwater. Groundwater modeling and field data submitted in response to this request, along with U.S. EPA's own data, showed that mercury could migrate from municipal landfills to contaminate drinking water supplies. ". . . [Actual] site data from recent and on-going studies support the Agency's conclusion that mercury is present in significant concentrations in both leachate and groundwater at non-hazardous waste landfill sites, including municipal solid waste landfills, and has migrated off-site to drinking water sources (in some instances in concentrations exceeding Federal drinking water standards)."²⁷ Data compiled by DTSC corroborates U.S. EPA's findings; landfill leachate samples analyzed in four separate studies contained detectable mercury, sometimes in excess of federal primary drinking water standards.²⁸

In a review of data from California landfills in the Waste Management Unit Database System (WMUDS), mercury concentrations exceeded the MCL in three of 13 wells analyzed. The maximum concentrations were 0.004 mg/L in water sampled at the Tri-cities and Victorville landfills. One of five leachate samples analyzed contained mercury in excess of the MCL: a sample from the Zanker Road Landfill, which contained 0.0032 mg/L mercury.²⁹

In addition to concerns about the leaching of elemental and inorganic mercury from landfills, a recent study shows that methylmercury can be formed by bacteria in landfills, and can be directly emitted to air. Lindberg, *et al.*, report that various mercury species were detected in landfill gas from a Florida municipal landfill.³⁰ Total gaseous mercury was detected at concentrations in the $\mu\text{g}/\text{m}^3$ range, dimethylmercury was found in the ng/m^3 range, and methylmercury was detected in landfill gas condensate. The total gaseous mercury concentrations detected were "comparable to Hg levels in flue gas and the immediate downwind plume of coal-fired power plants . . ."³¹ The authors suggest that direct landfill emissions to air may account for methylmercury that has been detected in continental rainfall.

D. Mercury Contaminated Sites in California

1. CalSites Data

Since the mid-1980s, generators of mercury-containing waste have been subject to hazardous waste determination requirements. As discussed earlier, mercury-containing waste that meets hazardous waste identification criteria must be managed in accordance with storage, treatment, transportation, and disposal requirements designed to protect public health and environment. In spite of this extensive hazardous waste regulatory scheme, past and current human activities have led to unacceptable land contamination with mercury in some locations. Mercury-contaminated sites require assessment of the risks they pose to the public and the environment through all potential exposure routes. When a site's level of mercury contamination is found to pose significant risk, mitigation or cleanup is required.

DTSC's Site Mitigation Program maintains an automated database, which contains information on properties in California where hazardous substances have been released, or where the potential for a release exists. This database, referred to as "CalSites," is used primarily by DTSC staff as an informational tool to evaluate and track activities at properties that may have been affected by the release of hazardous substances. In April 2001, a search was completed for those sites where mercury was identified in CalSites as a known or suspected hazardous waste/substance. Eighty-one sites were identified in this search, ten of the eighty-one sites show DTSC actively working to remediate either in a lead role or in a support capacity.

It should be noted that the CalSites database should not be considered to be the sole database for identifying sites in California that contain mercury contamination.

2. Tailings Dumps

Past mining of cinnabar in California's Coast Range created mine-tailing dumps. These dumps contain significant amounts of exposed residual mercury. Tailings dumps contribute to environmental mercury loading two ways: they directly contaminate the land, and their mercury can leach and migrate, contributing to California's water mercury burden. The efficiency of "mercury recovery during retorting ranges from 90 to 95 percent, which results in calcine [tailings] that may contain from 5 to 10 percent of the mercury originally present in the ore."³² Sulfur in the piles of tailings reacts with oxygen and rainwater to form sulfuric acid, which readily dissolves mercury in the ore and carries it into creeks.³³ One study found that more than 80 percent of the dissolved mercury in Marsh Creek -- a small coast range creek -- could be traced to a single pile of exposed tailings at an abandoned mercury mine site.³⁴

III. Mercury in California's Air

Mercury and mercury compounds (mercury) found in California's air are the result of emissions from both natural and anthropogenic sources. The California Air Resources Board (ARB) is the state agency that maintains the emissions inventory for mercury in the air. It should be noted that emissions of mercury into the air are transitory and are

eventually deposited onto either land or water where they contribute to the mercury concentrations found in those environmental media.

Natural sources of mercury air emissions include volcanoes, wild fires, degassing from the earth's crust, and evaporation from the world's oceans.³⁵ Anthropogenic mercury emissions originate from a number of sources, including point and area-wide sources. Point sources emitting mercury include electric generation facilities, refineries, and cement manufacturers. The primary area-wide sources of mercury emissions are windblown dust and waste burning. Other anthropogenic sources of airborne mercury include the breakage of mercury-containing lamps and laboratories (research and analytical).

A. Ambient Air Concentrations of Mercury in California

California's median air mercury concentration is below the Limit of Detection of 3.0 ng/m³. Ambient air mercury concentration data for the past ten years can be accessed at the following ARB web site: www.arb.ca.gov/aqd/toxics/statepages/hgstate.html

The median ambient air concentrations reported at this web site do not reflect elevated air concentrations that may occur near stationary sources of mercury emissions.

B. California Air Toxics Programs

California's air toxics programs began in the late 1980's. Mercury has been a substance of interest to these programs since their inception. The most significant of these programs include the Toxics Air Contaminant Program, the Air Toxics "Hot Spots" Program, and the Children's Environmental Health Protection Program. We will discuss each of these programs and how mercury is included in each of them.

1. The Toxics Air Contaminant Program (AB 1807)

The ARB and OEHHA have identified mercury as a Toxic Air Contaminant (TAC). The process for identification of TACs was initiated by Assembly Bill 1807 (AB 1807, Tanner, 1983), also known as the "Toxic Air Contaminant Identification and Control Act." The bill requires the ARB and OEHHA to use criteria relating to "the risk of harm to public health, amount or potential amount of emissions, manner of, and exposure to, usage of the substance in California, persistence in the atmosphere, and ambient concentrations in the community" in the prioritization for the identification and control of air toxics. If a substance is identified as a TAC, the ARB staff ". . . reviews the emission sources of an identified TAC to determine if any regulatory action is necessary to reduce the risk."³⁶

The information generated by the TAC process that resulted in mercury (and mercury compounds) being designated as a TAC can be found at the following ARB web site: www.arb.ca.gov/toxics/tac/toctbl

2. The Air Toxics "Hot Spots" Program (AB 2588)

With mercury's designation as a TAC, it is a substance for which facility operators must

§ Note: a significant amount of the mercury emitted to the atmosphere from the earth's oceans and crust is re-emitted anthropogenic mercury that was previously deposited.

estimate and report emissions as required by the Air Toxics "Hot Spots" Information and Assessment Act (AB 2588, Connelly, 1987). AB 2588 requires stationary sources to report the routine emissions of a list of substances. The Air Toxics "Hot Spots" Program's (the Program) primary goals have been to collect data on the emissions of toxic substances, to identify facilities whose toxic emissions have localized effects, to determine the health risks posed by these emissions, and to notify local residents of these risks. The program was further refined by Senate Bill 1731 (SB 1731, Calderon, 1992) which amended the "Hot Spots" Act to require operators of facilities whose emissions pose significant risks to reduce these risks until they are no longer significant.³⁷

Facility operators have been reporting Air Toxics "Hot Spots" emission inventory data to the ARB since 1989. Not all facilities statewide are subject to the Program. The ARB works closely with the local air pollution control districts and air quality management districts (the districts) to ensure that facilities that could potentially pose a risk to the quality of life of the local residents are required to submit emission inventories and to evaluate these potential risks. The mercury emissions estimates collected to meet the requirements of the Air Toxics "Hot Spots" Program serve as the ARB's statewide point source data for mercury air emissions.

3. The Children's Environmental Health Protection Program (SB 25)

Children can sometimes be more at risk than adults from the harmful health effects of air pollution. To provide further protection to children, the ARB is implementing a number of activities to evaluate and reduce those health risks. Senate Bill 25 (Escutia, 1999) established specific requirements to examine the impacts of air pollution on children's health. The ARB's efforts include:

- Review of ambient air quality standards to determine whether the standards adequately protect the health of the public including children,
- Revision of those standards found to be inadequate,
- Expansion of monitoring for air pollutants to assess the monitoring network's ability to measure children's exposure to air pollution, and
- Identification and control of TACs to which children may be especially sensitive (the most significant of those TACs will be determined by OEHHA).

Mercury is one of the substances that are being monitored in the Children's Environmental Health Protection Program. The scientific review panel working with OEHHA to prioritize the most significant substances has placed mercury in the second tier of concern. This decision was based on several factors including mercury's relatively high neurological and developmental toxicity, but its low ambient levels in California. Additional information about the Children's Environmental Health Protection Program can be obtained by visiting the following web site:

www.arb.ca.gov/ch/ceh/ceh.htm

C. Mercury Health Data Associated with Air Exposures

As part of the process of evaluating risks under the Air Toxics "Hot Spots" Program, ARB and OEHHA have approved Reference Exposure Levels (RELs) for some of the TACs, for use in health risk assessments. RELs have been developed for inorganic mercury and compounds, and for organic mercury and compounds. OEHHA defines an REL as the "... concentration level at or below which no adverse health effects are anticipated for a specified exposure duration"38 The approved REL values for mercury and mercury compounds for chronic inhalation, chronic oral, and acute inhalation are summarized in Table 1-1.

Table 1-1: OEHHA/ARB Approved Risk Assessment Health Values³⁹

NON-CANCER EFFECT (UNITS)	MERCURY AND COMPOUNDS (INORGANIC)	MERCURIC CHLORIDE	MERCURY AND COMPOUNDS (ORGANIC)**
Acute Inhalation ($\mu\text{g}/\text{m}^3$)	1.8	1.8	NA††
Chronic Inhalation ($\mu\text{g}/\text{m}^3$)	0.09	0.09	1.0
Chronic Oral (mg/kg/day)	0.0003	0.0003	NA††

D. Occupational Exposure Standards

It is relevant to note when discussing air standards and acceptable inhalations risks that a variety of industrial hygiene standards have been established for several different forms of mercury to protect occupationally exposed workers from mercury's toxic effects. Some of these standards are enforceable, while others are advisory in nature. Table 1-2 summarizes some of the existing standards metallic, inorganic, and organic mercury.

** Values also apply to methylmercury
 †† NA = None adopted

Table 1-2: Industrial Hygiene Limits for Occupational Exposure (mg/m³) Mercury Inhalation

MERCURY FORM	CAL - OSHA PEL ^{††}			ACGIH TLV ^{§§}			NIOSH ^{***} REL/IDLH			
	PEL ^{†††}	STEL ^{†††}	C ^{§§§}	TLV ^{****}	STEL	C	REL ^{††††}	STEL	C	IDLH ^{††††}
Mercury Vapor	0.05	--	0.1	0.025	—	—	0.05	—	—	10
Alkyl Mercury (organo)	0.01	0.03	0.04	0.01	0.03	—	0.01	0.03	—	2
Aryl and inorganic Compounds	—	—	0.1	0.1	—	—	N/A	N/A	N/A	N/A

It should be noted that the industrial hygiene occupational exposure levels to all forms of mercury are significantly higher than the RELs established by OEHHA. For example, the federal Occupational Safety and Health Administration's (OSHA's) PEL for mercury vapor of 0.05 mg/m³ or 50 µg/m³, is approximately 500 times higher than the REL for chronic inhalation, which is 0.09 µg/m³. These differences may be attributable to differences in the risk assessment methodology and default assumptions that were used to derive the respective values.

E. Air Emissions

The ARB stores statewide air emissions data in the California Emission Inventory Development and Reporting System (CEIDARS). CEIDARS contains emissions information for criteria pollutants (oxides of nitrogen, total organic gases, particulate matter, etc.) and for toxic substances. These data are gathered for stationary, area-wide, on-road mobile, off-road mobile, and natural sources. Inventories of emissions to air are revised on an annual basis to reflect the addition or deletion of sources, revised emission estimation methodologies, and revised speciation profiles. The emission estimates of mercury cited in this report are associated with the emission inventory for

†† **Cal-OSHA** – California Occupational Safety and Health Administration.

§§ **ACGIH** - American Conference of Governmental and Industrial Hygienists.

*** **NIOSH** - National Institute of Occupational Safety and Health. Mercury vapor includes both aryl and inorganic mercury.

††† **PEL** - Permissible Exposure Limit. The maximum permitted 8-hour time-weighted average concentration of an airborne contaminant.

††† **STEL** - Short-term exposure limit. A 15-min time-weighted-average exposure that should not be exceeded at any time during a workday even if the 8-hour time-weighted-average is within the threshold limit value.

§§§ **C** – Ceiling. These values should not be exceeded at any time.

**** **TLV** – Threshold Limit Value. The time-weighted average concentration for a conventional 8-hour workday and a 40-hour workweek, to which it is believed that nearly all workers may be repeatedly exposed, day after day, with our adverse effect.

†††† **REL** – Recommended Exposure Levels. These are time-weighted averages (TWA) concentrations for up to a 10-hour workday during a 40-hour workweek.

†††† **IDLH** - Immediately Dangerous to Life or Health. The maximum environmental concentration of a contaminant from which one could escape within 30 min without any escape-impairing symptoms or irreversible health effects.

the 1996 calendar year.

The stationary sources in CEIDARS are categorized as point sources and aggregated point sources. Generally speaking, a point source is a facility that emits greater than ten tons per year of one or more of the criteria pollutants. The aggregated point sources are those smaller facilities that have significantly similar emissions and a relatively small number of processes associated with them. Gasoline service stations and dry cleaners are examples of aggregated point sources. Emissions estimates for the vast majority of these facilities are developed by the facility operator under the auspices of the Air Toxics "Hot Spots" Program, but there are cases where the districts will develop the emission estimates for the facility operators. The mercury compounds facility operators are required to estimate and report include mercury, mercuric chloride, and methyl mercury.

Area-wide sources are estimated by the ARB and include the very small individual sources (residential combustion sources are an example) and the widely distributed sources that cannot be tied to a single location (consumer products, for example). Emission estimates of toxic substances, such as mercury, are developed by speciating criteria pollutant emissions associated with these sources.

Emissions from on-road and off-road mobile sources are estimated using California-specific models developed by the ARB. The sources included in the on-road model include cars, trucks, and buses. The sources included in the off-road model include aircraft, recreational equipment, and agricultural equipment. Emissions of toxic substances from mobile sources are developed by speciating criteria pollutant emissions associated with these sources.

Sources of emissions from natural sources include biogenics and wild fires. Emissions of toxic substances from natural sources are developed by speciating criteria pollutant emissions associated with these sources.

Specific mercury emission estimates from each of these source types will be discussed in detail in Sections 3 and 5 of this report.

IV. Water Mercury Burden

A. Background/Ambient Water Quality

Open ocean concentrations of dissolved mercury have been measured between 0.5 ng/l and 3.0 ng/L, while coastal concentrations were measured higher, ranging from 2 to 15 ng/L.⁴⁰ Both concentration ranges are well below the recommended ambient water quality criterion of 50 ng/L. However, ambient concentrations in some water bodies exceed this criterion. For example, the San Francisco Bay RWQCB reports preliminary mercury concentrations ranging from 2 to greater than 100 ng/L in the San Francisco Bay⁴¹.

B. Standards

1. Types of Water Quality Goals

There are many water quality criteria and goals that are designed to protect specific beneficial uses of water. These water quality goals can be used to interpret narrative water quality objectives. Table 1-3 summarizes the main water quality goals that are discussed in this document. The Reference section at the end of this report lists the sources of these limits, including Internet addresses, where available.

Table 1-3: Summary of Water Quality Goals in California⁴²

Water Quality Goal	Agency	Law	Meaning
Maximum Contaminant Levels (MCLs)	California Department of Health Services (DHS)	California Safe Drinking Water Act.	MCLs are set a level as close as is technically and economically feasible to the public health goal (PHG) (see below), placing primary emphasis on the protection of public health. <u>Carcinogens</u> : often set at or near the level of up to one excess case per million people per 70-year lifetime exposure, but may be less restrictive because of technical and economic feasibility. <u>Non-carcinogens</u> : set at level that would pose no adverse health effects.
Maximum Contaminant Level Goals (MCL Goals or MCLGs)	U.S. EPA	National Primary Drinking Water Regulations	<u>Carcinogens</u> : zero. <u>Non-carcinogens</u> : levels posing no risk of adverse health effects.
Public Health Goals (PHGs)	OEHHA	California Safe Drinking Water Act of 1996	Levels of contaminants in drinking water that would pose no significant health risk to individuals consuming the water on a daily basis over a lifetime.
State Action Levels	DHS		<u>Carcinogens</u> : one excess case per million people for a lifetime exposure <u>Non-carcinogens</u> : a level that would pose no adverse health effects
California Environmental Protection Agency (Cal/EPA) Cancer Potency Factors	OEHHA		Cancer potency factors for inhalation and oral exposures to many chemicals.
Integrated Risk Information System (IRIS)	U.S.EPA Office of Research and Development National Center for Environmental Assessment		Reference doses (RfDs): calculated safe exposure levels with respect to non-cancer health effects. RfDs may be converted into concentrations in drinking water (mg/L or µg/L) using standard exposure assumptions.
Drinking Water Health Advisories and Water Quality Advisories	U.S. EPA		Advisories for short-term (1-day exposure or less or 10-day exposure or less), long-term (7-year exposure

Water Quality Goal	Agency	Law	Meaning
			or less), and lifetime human exposures through drinking water.
Suggested No-Adverse-Response Levels (SNARLs)	National Academy of Sciences (NAS)		Published in the nine volumes of Drinking Water and Health (1977 to 1989).
Proposition 65 Regulatory Levels	OEHHA	California Safe Drinking Water and Toxic Enforcement Act of 1986	Requires notification prior to exposing persons to listed carcinogens or reproductive toxins, and prohibits discharges to sources of drinking water. Warnings are not required and discharges are not prohibited if: for <u>carcinogens</u> , risks are at one per 100,000 lifetime risk or lower; <u>reproductive toxins</u> , exposures are less than 1/1,000 of the no observable adverse effect level.
National Ambient Water Quality Criteria	U.S. EPA	Section 304(a) of the Clean Water Act	Provide guidance to states in adopting water quality standards. Concentrations based on exposure from drinking water and consuming aquatic organisms (fish and shellfish) that live in the water.
California Toxics Rule (CTR) Criteria	U.S. EPA	Federal Clean Water Act	U.S. EPA-promulgated water quality criteria for priority toxic pollutants for California's inland surface waters and enclosed bays and estuaries.

Some of these goals/limits have been established for mercury. These are summarized in Table 1-4, below.

Table 1-4: Summary of State and Federal Water Quality Standards for Mercury
Units are micrograms per liter (µg/L)

Inorganic Constituent			INORGANIC MERCURY	MERCURIC CHLORIDE
Drinking Water Standards (Calif. And Federal) MCLs	CA DHS Primary MCL		2	
	U.S. EPA Primary MCL		2	
	U.S. EPA MCL Goal		2	
OEHHA Public Health Goal (PGH) in Drinking Water			1.2	
U.S. EPA IRIS RFD as a Drinking Water Level				0.2
U.S. EPA SNARL for non-cancer Toxicity			2	
California Prop 65 Level as a Drinking Water Level			R*****	R*****
U.S. EPA National Recommended Ambient Water Quality Criteria	Non-cancer Effects—Drinking Water Sources (water and organisms)		0.050	
	Non-cancer Effects—Other Waters (aquatic organism consumption only)		0.051	
	Freshwater Aquatic Life Protection—Recommended Criteria	Continuous concentration (4-day Average)		0.77
Maximum Concentration (1-hour Average)		1.4		
California Toxics Rule (U.S. EPA)	Inland Surface Waters	Human Health (30-day average)	Drinking Water Sources (consumption of water and organisms)	0.05
			Other Waters (aquatic organism consumption only)	0.051
	Enclosed Bays and Estuaries	Human Health (30-day average) aquatic organism consumption only		0.051
California Ocean Plan Numerical Water Quality Objectives	Marine Life Aquatic Protection		6-month Median	0.04
			Daily Maximum	0.16
			Instantaneous Maximum	0.4
U.S. EPA National Recommended Ambient Water Quality Criteria—Saltwater Aquatic Life Protection	Recommended Criteria		Continuous Concentration (4-day average)	0.94
			Maximum Concentration (1-hour average)	1.8

The wildlife values as they appear in the Federal Register should be included here.

§§§§ From: California Regional Water Quality Control Board, Central Valley Region: *A Compilation Of Water Quality Goals*
***** Reproductive Toxin

2. Total Maximum Daily Loads (TMDL) for Mercury

The Federal Clean Water Act requires that California identify water bodies that do not meet water quality standards and develop total maximum daily pollutant loads for those water bodies. A TMDL represents the total loading rate of a pollutant that a water body can receive and still meet applicable water quality standards. Once a TMDL for a particular pollutant has been established, the load is allocated to all sources in the watershed, point and non-point, which must implement control measures as needed to reduce their discharges to the levels allocated to them. The San Francisco Bay RWQCB, in its TMDL Report, has proposed a sediment mercury target of 0.20 µg/g, and targets for methylmercury in bay fish that are 50 percent below current levels.⁴³

C. Water Mercury Sources

Mercury can enter impacted water bodies like the San Francisco Bay estuary from a variety of sources. Because of mercury's tendency to adsorb to particulates, the remobilization of contaminated sediments can be a significant source of mercury loading. The San Francisco Bay RWQCB has identified remobilized sediments from the Central Valley as the largest source of mercury loading in the San Francisco Bay.⁴⁴ The next largest mercury input identified by the RWQCB is the remobilization of contaminated sediments within the Bay that are gradually being eroded away.⁴⁵ Other important sources are watersheds within the San Francisco Bay Estuary, direct discharge of mercury-containing wastewater, and direct atmospheric deposition.⁴⁶ The relative contributions of these sources may differ in other impacted water bodies.

V. Public Health / Environmental Issues

Many regulatory efforts are already underway to reduce environmental mercury loading. They include management requirements for hazardous waste, mandates for the reduction of air emissions from stationary sources, point source controls on wastewater discharges, occupational exposure limits for mercury, and bans on the use of mercury in consumer products. Additionally, efforts are ongoing to mitigate and clean up contaminated sites. These activities are designed to reduce the potential exposure of humans, wildlife and the environment and the risks that such exposures entail.

The risks posed to humans and wildlife from environmental mercury exposure can be estimated through a process known as risk assessment. OSHA and industrial hygiene advisory groups also use a risk assessment process, which is specific to a workplace exposure setting, to determine the occupational exposure limits. Risk assessment involves the evaluation of potential exposure routes to the sensitive receptor (human or wildlife). The concentrations of a substance that can be assimilated by the sensitive receptor through all potential exposure routes are determined, and are compared to a reference dose. (A reference dose is one that is considered acceptable over the receptor's lifetime.) The specific details of the risk assessment process are not within this scope of this report.

A. Methylmercury in Fish / Consumption Advisories

Although regulatory standards limiting releases of mercury into the environment are in place, mercury's ability to move from air and soil into water continues to pose a public health risk. This risk is due to methylmercury's propensity to bioaccumulate in fish and human consumption of methylmercury contaminated sport fish. Using reference doses and complex risk assessment calculations, OEHHA has determined that mercury fish advisories are necessary in California's recreational waters.

Currently, there are OEHHA advisories against the consumption of any fish from the Guadalupe, Caldero, and Almaden Reservoirs, as well as the Guadalupe River and Guadalupe and Alamitos Creeks as a result of mercury contamination originating from nearby abandoned mines.⁴⁷ OEHHA has issued several other fish consumption advisories due in part or entirely to mercury contamination.⁴⁸ These advisories specify maximum consumption limits for specific fish species and sizes. These are summarized in Table 1-5.

Table 1-5: Sport Fish Consumption Advisories for Mercury Contaminated Water Bodies, 1999

Affected Water Body	Consumption Limits General Population	Fish Species
Clear Lake	<p><u>Adults</u>: ranges from 1 lb. to 10 lbs. per month, depending on species and size.</p> <p><u>Children aged 6 to 15</u>: half the maximum amounts recommended for adults, ranging from 0.5 lb. to 5 lbs. per month, depending on species and size.</p> <p><u>Pregnant/nursing mothers, children under 6</u>: No consumption.</p>	Bass (largemouth and smallmouth), catfish (white and channel), trout (rainbow), brown bullhead, Sacramento blackfish, crappie, hitch
Lake Berryessa	<p><u>Adults</u>: ranges from 1 lb. to 10 lbs. per month, depending on species and size.</p> <p><u>Children aged 6 to 15</u>: half the maximum amounts recommended for adults, ranging from 0.5 lb. to 5 lbs. per month, depending on species and size.</p> <p><u>Pregnant/nursing mothers, children under 6</u>: No consumption.</p>	Bass (largemouth and smallmouth), catfish (white and channel), trout (rainbow)
San Francisco Bay/Delta (interim)	<p><u>Adults</u>: no more than two 8-oz. meals per month. No striped bass over 35 inches.</p> <p><u>Pregnant/nursing mothers, children under 6</u>: no more than one 8-oz. meal per month. No striped bass over 27 inches or shark over 24 inches.</p> <p><u>Everyone</u>: no croakers, gobies, or shellfish from the Richmond Harbor Channel area.</p>	Sport fish, including sturgeon and striped bass from the delta
Lake Hermann	<p><u>Adults</u>: no more than 1 lb. largemouth bass per month.</p> <p><u>Children aged 6 to 15</u>: no more than 8 oz. largemouth bass.</p> <p><u>Pregnant/nursing mothers, children under 6</u>: No consumption.</p>	Largemouth bass Largemouth bass Any fish
Guadalupe Reservoir	No consumption.	Any fish
Calero Reservoir	No consumption.	Any fish
Almaden Reservoir	No consumption.	Any fish

Affected Water Body	Consumption Limits General Population	Fish Species
Guadalupe River	No consumption.	Any fish
Guadalupe Creek	No consumption.	Any fish
Alamitos Creek	No consumption.	Any fish

B. Mercury Contaminated Sites

Mercury-contaminated sites listed in the CalSites database were previously discussed. Mercury may be the only hazardous contaminant present at a contaminated site, or it may be one of many chemicals of concern. After a site is fully characterized, a risk assessment is performed. Typically during the site characterization process, public access to a contaminated site is restricted, in order to reduce any potential exposure of the public to the chemicals of concern. If necessary, cleanup activities or mitigation measures are performed on the contaminated site.

In spite of the fact that these measures are taken to assess and clean up land contamination, unintentional contamination of land with mercury continues to be an issue. Contamination may occur through disposal of non-hazardous mercury-containing waste in Class III landfills, or through illegal garbage dumping in rural areas. Because of the persistence and bioaccumulative properties of mercury, nonhazardous waste that contains mercury may add to the current risk to public health and environment.

C. Nontraditional Sources of Mercury

Some activities that lead to human exposure to mercury occur outside of the workplace, and fall outside of the California OSHA's regulatory authority. These include recreational and hobby activities. Although measures are taken to educate the public of the dangers of mercury, these activities are not formally regulated in California, although they may pose risks to the public. They include recreational gold mining, where recovered gold is often found amalgamated with mercury. Some recreational gold miners refine gold at their homes, exposing themselves to mercury in the process, as well as emitting mercury to the air. Waste liquid mercury collected in the course of recreational gold recovery is either disposed in an environmentally sound manner through household hazardous waste collections, or disposed onto land or in the sewer via the toilet, causing an additional mercury burden to the State's waters.

Section 1 Key Points:

- Mercury is ubiquitous in the environment due to its natural occurrence and its widespread current and historical use.
- Mercury is a contaminant of special concern because of its toxicity, persistence, environmental mobility, and ability to bioaccumulate.
- Mercury's health and environmental hazards have led to the development of numerous regulatory standards for mercury in waste, air, and water.
- Standards for occupational exposure airborne to mercury have also been adopted, due to its health hazards.
- In spite of the existing regulatory standards, California's environment continues to be contaminated with mercury.
- Airborne mercury is a concern because it is eventually deposited on land and water.
- Mercury is contained in waste that is classified as non-hazardous under current regulatory criteria.
- The disposal of non-hazardous products contributes a significant amount of mercury to municipal landfills.
- Mercury can dissolve in landfill leachate and potentially contaminate the State's waters.
- Aquatic mercury is converted to a very toxic and bioaccumulative form, methylmercury, by certain bacteria.
- Mercury land contamination at a number of sites in California has made cleanup or other mitigation activities necessary.
- Some of California's water bodies exceed water quality standards for mercury. The Federal Clean Water Act requires that total maximum daily loads be developed for mercury in these water bodies.
- California's Office of Environmental Health Hazard Assessment has advised the public to restrict or eliminate its consumption of specific sport fish from several water bodies, due to elevated levels of methylmercury in the fish.

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Section 2: Mercury's Chemistry and Toxicology--Human and Environmental Hazards

I. Introduction

Mercury's health and environmental hazards stem from its toxicity and its mobility in the environment. As discussed briefly in Section 1, each form of mercury has distinct chemical and physical properties and toxicology. This section provides an overview of the chemistry of the three most environmentally important forms, their role in the global mercury cycle, and their toxicology. These discussions provide context for the concerns about the hazards of mercury in California's environment and the adequacy of the current efforts to control human contributions to the State's mercury problem.

II. Physical and Chemical Properties of Mercury and Mercury Compounds

A. Melting Point, Volatility

Mercury can exist in three oxidation states: Hg⁰ (elemental or metallic), Hg¹⁺ (mercurous), and Hg²⁺ (mercuric). The physical and chemical properties of these species differ significantly, as can be seen in Table 2-1, which compares some important properties of elemental mercury, mercuric chloride (an environmentally significant inorganic form), and methylmercury (an environmentally significant organic form).

Table 2-1: Physical and Chemical Properties of Selected Mercury Species¹

Mercury Species	Elemental Mercury	Mercuric Chloride	Methylmercury ^{*2}
Formula	Hg	HgCl ₂	CH ₃ HgCl
Atomic/Molecular Weight	200.59	271.52	251.10
Density	13.53 @ 25° C	5.4 @ 25° C	3.18 @ 20° C
Vapor Pressure	0.002 mm Hg @ 25° C ³		0.0085 mm Hg @ 25° C ⁴
Melting Point (°C) ⁵	-38.87°	276° ⁶	170°
Boiling Point (°C) ⁷	356.9°	302° ⁸	No data
Solubility (grams per liter) ⁹	5.6 x 10 ⁻⁵ @ 25°C	69 @ 20°C	0.100 @ 21° C

Metallic mercury is almost unique among metals in that it is a liquid at room temperature.†¹⁰ This fact, along with its relatively high vapor pressure, accounts for the wide dispersal of mercury in the environment.

B. Covalent Bonding with Carbon

Another important property of mercury is its ability to form covalent bonds with carbon. Compounds that consist of an organic functional group covalently bonded to a metal are known as organometallic compounds. They are often highly toxic, and organomercurics are especially so.

C. Important Mercury Compounds

The best known organomercuric is the very toxic compound methylmercury, which

* "Because methylmercury exists as a free ion only in minute quantities (Prager, 1997), the chemical and physical data . . . are for the chloride salt." (U.S. EPA Water Quality Criterion for the Protection of Human Health: Methylmercury, 2001.)

† Gallium and Cesium are the only other metals that are liquids at room temperature.

typically occurs as the salts methylmercuric chloride (CH_3HgCl) and methylmercuric hydroxide (CH_3HgOH).¹¹ The most environmentally significant inorganic mercury salts are mercuric chloride (HgCl_2), mercuric hydroxide [$\text{Hg}(\text{OH})_2$], and mercuric sulfide or cinnabar (HgS).

D. Solubility of Mercury and Mercury Compounds

The water solubility of the various forms of mercury varies widely. Least soluble is metallic mercury, at 5.6×10^{-5} grams per liter (at 25°C). At 0.100 grams per liter (at 21°C) methylmercury is nearly 2,000 times more water-soluble; still more soluble is mercuric chloride, at 69 grams per liter (at 20°C).¹²

E. Unique Properties

Mercury is unique, in that it:

- Is a liquid at room temperature;
- Forms **ef** covalent bonds; and
- Has a relatively high vapor pressure.

III. The Global Mercury Cycle - Mercury Environmental Fate and Transport

A. The Global Mercury Cycle (Environmental Mercury Fluxes)

The global mercury cycle is described in terms of the flux (movement) of mercury between environmental media. The mercury flux at a given location includes global, regional, and local contributions. Regional and local mercury fluxes vary widely, so it is difficult to generalize about them, but the global cycle (and the contribution of anthropogenic inputs) is well characterized. Studies by Nriagu (1979) and Fitzgerald (1994), summarized in Table 2-2, both conclude that the vast majority of the world's environmental mercury is found in ocean sediments.

Table 2-2: Estimated Mercury Content of Environmental Media – Worldwide^{13,14}

	Grams (g) Nriagu (1979)	Grams (g) Fitzgerald (1994)
Ocean Sediments	10^{17}	
Ocean Waters	10^{13}	
Freshwater Sediments	10^{13}	
Biosphere	10^{11}	
Atmosphere	10^8	5×10^9
Fresh Water	10^7	

Some authors have estimated the mercury concentrations in the various environmental media prior to industrialization. However, such estimates are difficult to make, because the current environmental mercury that is of anthropogenic origin is indistinguishable from that which was naturally emitted. The consensus in these studies is that between 40 and 75 percent of the mercury emitted to the atmosphere, worldwide, is of anthropogenic origin.¹⁵ U.S. EPA believes that more study is needed in order to make it

possible to distinguish natural mercury fluxes from fluxes of re-emitted anthropogenic mercury.

B. Fate and Transport of Mercury

1. Atmospheric

a. Deposition of Atmospheric Mercury

All atmospheric fluxes of elemental mercury, worldwide, contribute to a global pool of atmospheric mercury as mercury readily evaporates and is transported in air. Recent monitoring of atmospheric mercury levels show that the world's atmospheric burden has increased between two- and five-fold in industrialized times. Studies of Swedish lake sediments, Upper Midwest lakes and peat cores, and remote Alaskan lakes have corroborated these measurements.¹⁶ Much of the mercury emitted to the atmosphere from the oceans is recycled mercury of anthropogenic origin. One study estimates that only 20 to 30 percent of ocean-emitted mercury is of natural origin.¹⁷ A similarly large percentage of terrestrial mercury emissions may be remobilization of anthropogenic mercury.¹⁸

Before it is ultimately deposited on land or water, either through atmospheric precipitation (wet deposition) or through atmospheric particulate (dry deposition), most atmospheric elemental mercury undergoes oxidization. U.S. EPA mentions two mechanisms for atmospheric oxidation. Most important of these is the oxidation of gaseous elemental mercury to aqueous and particulate-associated divalent mercury (Hg^{+2}) in cloud water. Another (less significant) process mentioned by U.S. EPA is the ozone-mediated oxidation of metallic mercury to divalent mercury, which is then dry-deposited on land or water.¹⁹

Gas-phase divalent mercury is both reactive and soluble in water. Consequently, this form is "rapidly and efficiently removed by both dry and wet deposition . . ." from the atmosphere. Elemental mercury, on the other hand, is relatively insoluble in water and has a higher vapor pressure; unlike the divalent form, it is "not thought to be susceptible to any major process of direct deposition."²⁰ U.S. EPA cites a number of studies that describe a minor mechanism for direct deposition of elemental mercury: uptake by the leaves of plants. The studies show that elemental mercury vapor can be taken up by leaves in forest canopies. One study (Hanson, et al., 1994) found that, while such leaf uptake can occur, the net flux of mercury from plants to air is generally higher than that from air to plants. It found that plants can be a net sink for elemental mercury vapor when ambient air mercury concentrations are sufficiently high.²¹

b. Half-life of Mercury in the Atmosphere

Some atmospheric mercury is deposited on land or water relatively near to the emission source, while some enters the global atmospheric mercury cycle, where it is transported to the remotest regions of the earth. The U.S. EPA 1997 Study states that, on average, emitted elemental mercury resides in the atmosphere for one year. By contrast, divalent mercury is deposited relatively quickly, with a residence time as short as a few hours and several months. Consequently, elemental mercury that is emitted to air is distributed worldwide before it is ultimately deposited on land or water, while atmospheric divalent mercury is mostly deposited relatively close to the emission source.²² Porcella, et al. found that mercuric mercury associated with fine particulates

may, like metallic mercury, persist in the atmosphere for up to one year.²³ Because emitted elemental mercury generally persists in the atmosphere for much longer than the oxidized species, global transport and deposition of this form constitute by far the most significant atmospheric mercury flux.²⁴

Combustion and incineration are important categories of atmospheric mercury emissions. Stack emissions contain both oxidized and reduced (elemental) mercury. U.S. EPA states that gaseous emissions are thought to contain both forms; while in particulate emissions (soot), oxidized mercury predominates.²⁵

2. Terrestrial

Of the environmental media, mercury is least mobile in soil, which “results in soil acting as a large reservoir for anthropogenic mercury emissions.”²⁶ U.S. EPA states that divalent mercury compounds tend to form immobile complexes with organic matter and minerals in soil. However, it can form soluble complexes with organic ligands and subsequently dissolve in runoff. The current consensus, according to U.S. EPA, is that the rate of deposition of atmospheric mercury on soil greatly exceeds the rate of leaching of mercury from soil.²⁷ “Mercury that has accumulated in soils may continued to be released to surface waters and other media for long periods of time, possibly hundreds of years.”²⁸

Although mercury is less mobile in soil than in water and air, terrestrial mercury can migrate. As noted in Section 1, leaching of mercury from municipal landfills is noted in U.S. EPA's Universal Waste Landfill Rule proposal, in data compiled by DTSC, and in the SWRCB's Waste Management Unit Database System.

Wildlife are often attracted to landfills, and small mammals and birds cannot be excluded by fences designed to keep out humans and large mammals. Wildlife may therefore be more at risk from mercury-containing landfill leachate.

1. Bruner, M.A., et al., *Ground and surface water developmental toxicity at a municipal landfill: Description and weather-related variation*. *Ecotoxicology And Environmental Safety*, 1998. **39**(3): p. 215-226.

1. Robinson, G.R. and S.N. Handel, *Forest Restoration on a Closed Landfill Rapid Addition of New Species by Bird Dispersal*, in *Human urbanization & Conservation*, D. Ehrenfeld, Editor. 1995. p. 222-229.

3. Weber, J., *Mercury, a Hazardous Waste Problem*. *J. Environ. Health*, 1983. **54**(6): p. 284-287.

4. Bruner, M.A., et al., *Ground and surface water developmental toxicity at a municipal landfill: Description and weather-related variation*. *Ecotoxicology And Environmental Safety*, 1998. **39**(3): p. 215-226.

5. Gabrey, S.W., *Bird and small mammal abundance at four types of waste-management facilities in northeast Ohio*. *Landscape And Urban Planning*, 1997. **37**(3-4): p. 223-233.

3. Fresh Waters

Methylmercury and divalent mercury can enter freshwater environments by several routes: via wet or dry atmospheric deposition, via runoff from land, and via leaching in groundwater.²⁹ Once it enters the freshwater environment, divalent mercury can form immobile complexes by the same processes as occur on land.³⁰ In aquatic environments, both methylmercury and inorganic divalent mercury preferentially partition to soil, sediment, and suspended matter (i.e., dissolved mercury concentration is far lower than the concentration in soil, sediment, and suspended matter).³¹ Most mercury in the water column is bound to dissolved organic carbon or bound to suspended particles.³² According to U.S. EPA, divalent mercury is reduced to the elemental species in the freshwater environment and may subsequently be removed from the water column by volatilization. Studies cited by Mason, et al., show that most such reduction is biologically mediated.³³ However, most of the mercury in the water column is removed not by reduction to the elemental species, but by sedimentation of the particles to which divalent mercury and methylmercury are bound.³⁴

The methylation of mercury in aquatic environment is critically important in the global mercury cycle, because methylmercury is an especially bioavailable form of the metal.³⁵ The biological process by which methylmercury is formed, in conjunction with bioaccumulation and biomagnification of methylmercury in animals that live in contaminated waters and animals that prey upon them, are important components of the biogeochemical mercury cycle. U.S. EPA cites studies that show that methylation can occur both in the water column and in sediments, by both biological and abiotic processes.³⁶ Jones and Slotton identify several factors that affect the rate of mercury methylation in aquatic sediments. These are summarized in Table 2-3.

***Add more on methylation here

Table 2-3: Environmental Influences on the Rate of Methylation of Aquatic Mercury³⁷

Environmental Factor	Effect On Mercury Methylation Rate
pH	Methylmercury is produced, transported, accumulated much more efficiently at lower pH.‡
Salinity	Increasing salinity decreases the amount of dissolved mercury, the rate of mercury methylation, and equilibrium methylmercury concentration.
Sulfate Concentration	Sulfate concentration affects the rate of mercury methylation; the maximum rate of methylation is seen when the sulfate concentration is between 200 and 500 mmol.
Oxygen concentration	Production of methylmercury is favored in anaerobic waters, as is its transfer to the food chain.

4. Marine Waters

A large percentage of the earth's mercury is found in oceanic waters and sediments (see Table 2-2). U.S. EPA states that atmospheric mercury, which is mainly in the elemental form, enters the world's oceans primarily by wet deposition.³⁸ As mentioned earlier, it is thought that elemental mercury is oxidized in the atmosphere. Oxidized

‡‡ The fact that California's waters have a naturally alkaline pH has mitigated the state's mercury problem somewhat.

mercury is more water-soluble and this property facilitates its deposition into water.³⁹

Marine mercury is transformed from one state to another by both biotic and abiotic chemical processes.⁴⁰ The U. S. EPA 1997 Study discusses two models of mercury's fate and transport in the ocean. One, developed by Fitzgerald and others, applies to the ocean as a whole; the other, developed by Cossa et al., applies to the waters at the margins of continents.⁴¹

a. Whole Ocean Model

In the model put forth by Fitzgerald, et al.⁴², reactive (e.g., divalent) mercury is first deposited on the ocean's surface. From there, it is transported downward with particles to the anoxic region below the thermocline (the boundary between the warmer, oxygen-rich waters of the surface and the colder, anoxic waters of the depths). As the particles descend, mercury is released and is methylated. Some of the methylmercury then moves to the upper, mixed layer, where it is taken up by organisms at the lowest levels of the food web. Some is reduced to the elemental form, by both biotic and abiotic processes, and is subsequently evaded from the water to the atmosphere. In coastal regions, the model assumes that mercury undergoes methylation in sediments and in the water column near the oxycline (defined as the "horizontal boundary layer in the water column, at which dissolved oxygen content changes sharply with depth"⁴³).

b. Continental Margin Model

The mercury mass balance model first developed by Cossa, et al. in 1996, identifies river sediments as the largest input of mercury to coastal waters. The model also assumes that coastal waters are subject to higher rates of atmospheric mercury deposition than those of the open ocean, primarily due to nearby emissions of reactive mercury. Another major flux to coastal waters identified in the model is transport of mercury from other parts of the oceans. Three fluxes of mercury from coastal waters are also identified: sedimentation, transport to the open ocean, and evasion to the atmosphere.

The Cossa, et al. model also describes the relative importance of the various methylmercury inputs to coastal waters. These are summarized in Table 2-4.

Table 2-4: Significant Methylmercury Inputs to the World's Coastal Waters⁴⁴

Input to Coastal Waters	MegaMoles per Year (Mmol/yr.)
Upwelling From Other Parts Of The Ocean	0.1 – 0.2
Atmospheric Deposition	0.02
River Systems	0.01
Sediments	0.001

Methylation and Uptake

U.S. EPA describes two marine food webs in which methylmercury bioaccumulation occurs: one in the sediments at the bottom of coastal waters, consisting of larger invertebrates, and one in the water column, made up of plankton.⁴⁵ The invertebrates in both of these communities take up methylmercury into their tissues from the surrounding environment.

However, marine organism may be at less risk from MeHg exposure, because of co-exposure to

Se, which is antagonistic to MeHg, therefore providing protection.

1. Cuvin-Aralar, M.L. and R.W. Furness, *Mercury and selenium interaction: A review*. *Ecotoxicol. Environ. Safe.*, 1991. **21**(3): p. 348-364.
2. Ganther, H.E., et al., *Selenium: Relation to decreased toxicity of methylmercury added to diets containing tuna*. *Science*, 1972. **Science**(175): p. 1122-1124.
3. Goede, A.A. and H.T. Wolterbeek, *Have high selenium concentrations in wading birds their origin in mercury?* *Science Of The Total Environment*, 1994. **144**(0): p. 247-253.
4. Koeman, J.H., et al., *Mercury and selenium in marine animals and birds*. *Sci. Tot. Environ.*, 1975. **3**: p. 279-87.
5. Nishikido, N., et al., *Maternal selenium deficiency enhances the fetolethal toxicity of methyl mercury*. *Toxicol Appl Pharmacol*, 1987. **88**(3): p. 322-8.
6. Siegel, B.Z., et al., *The protection of invertebrates, fish and vascular plants against inorganic mercury poisoning by sulfur and selenium derivatives*. *Arch. Environ. Contam. Toxicol.*, 1991. **20**: p. 241-246.
7. Sugiura, Y., et al., *Selenium protection against mercury toxicity. Binding of selenohydril-containing ligand*. *J. Am. Chem. Soc.*, 1976. **98**: p. 2339-2341.

As is the case in freshwater systems, mercury is believed to be methylated primary in anoxic sediments by sulfur-reducing bacteria. One study cited by U.S. EPA⁴⁶ found that a particular species of mussel assimilated particle-bound methylmercury more readily than particle-bound inorganic mercury. Dissolved methylmercury and inorganic mercury were both taken up more efficiently by the mussels than their particle-bound counterparts. However, the authors concluded that particle-bound methylmercury is the major source of the metal in the mussels, because of its much greater abundance in the coastal marine environment than the dissolved form. U.S. EPA cites other studies showing similar uptake mechanisms in other benthic organisms, and transfer of mercury to carnivorous animals that prey on them.⁴⁷

IV. Toxicology of Mercury and Mercury Compounds

A. Elemental Mercury⁴⁸

1. Toxicokinetics

Inhalation is the most important route of entry for elemental mercury. About 80 percent of inhaled elemental mercury is absorbed by the body. Once absorbed, the elemental form is distributed throughout the body. Airborne metallic mercury is also absorbed through the skin. The rate of dermal absorption increases with air concentration. The National Academy of Sciences (NAS) states that elemental mercury's average rate of absorption is 0.024 ng/cm³ for every 1 mg/m³ in air. The elemental form also "readily crosses the blood-brain and placental barriers," according to NAS. Ingested elemental mercury is poorly absorbed in the digestive tract,⁴⁹ and "the majority of the ingested dose is excreted in the feces."⁵⁰

Elemental mercury's half-life in blood is estimated by NAS to be 45 days, but "appears to increase with increasing dose." The metallic form can undergo biotransformation in the body, whereby it is oxidized to the mercuric (Hg²⁺) form. The metallic form leaves the body in exhaled air, perspiration, and saliva. Metallic mercury that has been biotransformed to the mercuric form is excreted in feces and urine.

2. Toxic Effects⁵¹

a. Carcinogenicity

The human epidemiological studies that U.S. EPA found in the preparation of the U.S. EPA 1997 Study have major limitations. While none of the studies show a correlation between human exposure to elemental mercury and increased cancer incidence, one shows such a correlation in animals injected with elemental mercury.

b. Neurotoxicity

Neurotoxic effects are elemental mercury's most sensitive toxicological endpoint, in U.S. EPA estimation. U.S. EPA identifies the following neurological symptoms of elemental mercury toxicity:

- Tremors, of the hands and other body parts
- Changeable emotional state, including irritability, extreme shyness, loss of confidence, and nervousness
- Insomnia
- Muscular weakness, atrophy, and twitching
- Headaches
- Sensory loss
- Hyperactive tendon reflexes
- Reduced nerve conduction velocities
- Memory loss
- Impaired cognitive function

c. Renal Toxicity

U.S. EPA states that toxic effects are seen in the kidneys at higher exposure concentrations than those required to produce neurotoxic effects.

d. Pulmonary Toxicity

U.S. EPA also states that toxic effects are seen in the lungs at higher exposure concentrations than those required to produce neurotoxic effects.

e. Reproductive Toxicity

U.S. EPA identified some studies suggesting that elemental mercury may cause reproductive toxicity. In two of these studies, behavioral changes were noted in rats that were exposed to elemental mercury *in utero* and around the time of birth.

f. Cardiovascular Toxicity⁵²

U.S. EPA identifies several manifestations of the cardiovascular toxicity of elemental mercury. It is unclear from the literature, according to U.S. EPA, whether elemental mercury directly causes toxicity to the heart, or whether the observed effects result from elemental mercury's neurotoxicity. The effects include:

- Tachycardia
- Elevated blood pressure
- Heart palpitations

3. Reference Exposure Standards

a. U.S. EPA Reference Doses

U.S. EPA has developed limits for exposure to hazardous substances, known the Reference Dose (RfD) and Reference Concentration (RfC). These terms are defined on the Internet web site for U.S. EPA's Integrated Risk Information System (IRIS) as follows.⁵³

RfC: An estimate (with uncertainty spanning perhaps an order of magnitude) of a continuous inhalation exposure to the human population (including sensitive subgroups) that is likely to be without an appreciable risk of deleterious effects during a lifetime. It can be derived from a No Observed Adverse Effects Level (NOAEL), Lowest Observed Adverse Effects Level (LOAEL), or benchmark concentration, with uncertainty factors generally applied to reflect limitations of the data used. Generally used in U.S. EPA 's noncancer health assessments.

RfD: An estimate (with uncertainty spanning perhaps an order of magnitude) of a daily oral exposure to the human population (including sensitive subgroups) that is likely to be without an appreciable risk of deleterious effects during a lifetime. It can be derived from a NOAEL, LOAEL, or benchmark dose, with uncertainty factors generally applied to reflect limitations of the data used. Generally used in U.S. EPA's noncancer health assessments.

Reference doses and concentrations are used in risk assessments to determine public health and environmental impacts through air, water and soil exposure routes through inhalation and ingestion. Table 2-5 summarizes the RfC that has been established for elemental mercury.

Table 2-5: Reference Doses (RfDs) and Reference Concentrations (RfCs) for Mercury, Elemental⁵⁴

SUBSTANCE NAME	EXPOSURE ROUTE	DURATION OF EXPOSURE	TEST SPECIES	RFC (MG/M ³)	RFD (MG/KG-DAY)
Mercury, Elemental	Inhalation	Chronic	Human occupational studies	0.0003	Not available at this time.

b. Agency for Toxic Substances and Disease Registry (ATSDR) Minimal Risk Levels (MRLs) for Hazardous Substances⁵⁵

In response to a mandate in the Superfund Amendments and Reauthorization Act (SARA) of 1990, ATSDR has developed MRLs for hazardous substances commonly found at facilities on the Comprehensive Environmental Response, Compensation and Liability Act (CERCLA) National Priorities List (NPL). An MRL is an estimate of the highest exposure to a hazardous substance that is not likely to pose significant health risks over a given period of exposure. Inhalation MRLs are stated units of parts per million (ppm) or milligrams per cubic meter (mg/m³). Oral MRLs are in units of milligrams per kilogram body weight per day (mg/kg/day). The MRL values established by ATSDR for metallic mercury are summarized in table 2-6.

Table 2-6: Minimal Risk Levels (MRLs) for Mercury, Metallic -- March 1996⁵⁶

SUBSTANCE	EXPOSURE	DURATION OF	TOXIC	MRL VALUE
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NAME	ROUTE	EXPOSURE	ENDPOINT	
Mercury, Metallic	Inhalation	Acute	Developmental	0.00002 mg/m ³
Mercury, Metallic	Inhalation	Chronic	Neurological	0.000014 mg/m ³

The number and range of health reference standards along with their corresponding low acceptable daily doses illustrate the toxic nature of mercury.

B. Mercuric Mercury

1. Toxicokinetics

Mercuric mercury can enter the body via inhalation, ingestion, or dermal exposure. Aerosols of mercuric mercury can be absorbed through the lungs, but NAS does not provide data for the efficiency of absorption by this route. NAS estimates that the efficiency of absorption of ingested divalent mercury is between 7 and 15 percent and that the efficiency of dermal absorption in guinea pigs is in the 2 to 3 percent range.

The divalent form tends to concentrate in the kidneys in adults, and the amount retained depends on the dose. In exposed newborns however, it does not concentrate in the kidneys, but rather is distributed throughout the body. Mercuric mercury, unlike the elemental form, does not easily cross the blood-brain or placental barriers. Any mercuric mercury that does cross the placenta can enter the brains of fetuses and neonates more readily than those of older children and adults, due to the incomplete formation of the blood-brain barrier. Mercuric mercury has a blood half-life of that ranges from 19.7 to 65.6 days, according to NAS.

NAS cites evidence that mercuric mercury can undergo biotransformation. They mention an experimental study in which elemental mercury vapor was found to be exhaled by rodents after they were orally administered mercuric mercury. NAS also states that, while mercuric mercury does not undergo methylation in body tissues, it is methylated by gastrointestinal microbes. The routes of excretion of the mercuric form are via urine, feces, saliva, bile, sweat, air, and breast milk.

2. Toxic Effects⁵⁷

a. Carcinogenicity

U. S. EPA identified no studies suggesting mercuric chloride is carcinogenic in humans. However, some studies in which rodents that were force-fed mercuric chloride showed increased incidence of certain tumors in exposed rats.

b. Renal Toxicity

The most sensitive toxic endpoint in humans exposed to inorganic mercury is autoimmune glomerulonephritis, according to U. S. EPA. This inflammation of the kidney results from the mercury-induced formation of antibodies to the basement membrane of the glomeruli.

c. Reproductive Toxicity

U.S. EPA found studies suggesting exposure to inorganic mercury salts may result in reproductive toxicity, but believes these studies are flawed.

3. Reference Exposure Standards

Tables 2-7 and 2-8, respectively, summarize the RfD and MRLs that have been established for mercuric chloride.

Table 2-7: Reference Doses (RfDs) and Reference Concentrations (RfCs) for Mercuric Chloride⁵⁸

SUBSTANCE NAME	EXPOSURE ROUTE	DURATION OF EXPOSURE	TEST SPECIES	RFC (MG/M ³)	RFD (MG/KG-DAY)
Mercuric Chloride	Oral	Chronic	Brown Norway rat	Not available at this time.	0.0003

Table 2-8: Minimal Risk Levels (MRLs) for Mercury, Inorganic -- March 1996⁵⁹

SUBSTANCE NAME	EXPOSURE ROUTE	DURATION OF EXPOSURE	TOXIC ENDPOINT	MRL VALUE
Mercury, Inorganic	Oral	Acute	Renal/Urinary	0.007 mg/kg/day
Mercury, Inorganic	Oral	Intermediate	Renal/Urinary	0.002 mg/kg/day

Discussion of reference doses and MRLs are found above in metallic mercury section. In contrast to metallic mercury, inorganic mercury's reference dose is based on the oral route of exposure rather than inhalation route. The exposure potential of these two forms of mercury differ in that the inhalation of metallic mercury is unlikely to occur outside an occupational setting. Furthermore, metallic mercury is poorly absorbed in the digestive tract, whereas inorganic mercury's rate of absorption is higher, as discussed above.

C. Methylmercury

1. Toxicokinetics

According to NAS, inhaled methylmercury vapors can be absorbed by the lungs. Methylmercury is also well absorbed in the gastrointestinal tract; humans absorb 95 percent of the methylmercury in fish they consume, according to NAS. In experiments with guinea pigs, 3 to 5 percent of dermally applied methylmercury was absorbed within 5 hours.

Up to 10 percent of absorbed methylmercury is distributed to the blood, and 90 percent of this 10 percent resides in red blood cells. Methylmercury is lipophilic and readily crosses the blood-brain and placental barriers.

Methylmercury's half-life in blood is estimated to be 50 days. Its blood half-life is reduced in lactating females. Methylmercury's half-life in the body is estimated to be from 70 to 80 days, depending on the species, strain, and sex of the experimental animal being studied, as well as the dose administered. It slowly undergoes biotransformation and is converted to the mercuric form by an unknown mechanism. Bile and feces are the important routes of methylmercury excretion, most of which is in the mercuric form.

The half-life of Hg in seabirds has been estimated to be about 60days (Monteiro, 1995).

Monteiro, L. R., A. J. Furness and A. J. del Novo. 1995. Mercury levels in seabirds from the Azores, mid-North Atlantic Ocean. Arch. Env. Contam. Toxicol. 28: 304 - 309.2.

Toxic effects⁶⁰

U. S. EPA notes that, in human and animal studies, there is often a delayed onset of the symptoms of methylmercury toxicity, which may be attributable to metabolic changes. For example, in the 1956 Minamata Bay incident, the victims were exposed to high levels of methylmercury, but did not exhibit signs or symptoms of mercury toxicity for several years.⁶¹

This delayed effect has been mirrored in avian species; in MeHg - exposed mallards, impairment was measurable into the second generations (see below under Reproductive Toxicity)

a. Carcinogenicity

U. S. EPA identified a number of epidemiological studies that analyzed the correlation between methylmercury exposure and human carcinogenesis. They are of the opinion that these studies were seriously flawed. However, evidence of carcinogenicity was seen in some rodent studies that U. S. EPA identified. Kidney tumors were observed in orally exposed mice, but only when other signs of severe nephrotoxicity were also observed.

b. Neurotoxicity

U. S. EPA identifies the nervous system as the “critical target for methylmercury toxicity.”⁶² Neurotoxic symptoms that occur in neonates are identified below, under the heading “Reproductive Toxicity”. In adults, methylmercury neurotoxicity is characterized by “multiple central nervous system effects.”⁶³ These include:

- Ataxia (impairment of voluntary muscle coordination)
- Paresthesia (tingling sensations)
- ***Wildlife:
- Inorganic Hg exerts its greatest effect on the kidneys whereas MeHg is a potent embryo and nervous system toxicant. Methylmercury readily penetrates the blood brain barrier in birds, as in mammals, producing brain lesions, spinal cord degeneration, and CNS dysfunctions. Symptoms of acute MeHg poisoning in birds include reduced food intake leading to weight loss, progressive weakness in wings and legs, difficulty flying, walking and standing and an inability to coordinate muscle movements (Scheuhammer, 1987)

Scheuhammer, A. M. 1987. The chronic toxicity of aluminum, cadmium, mercury, and lead in birds: a review. Environ. Pollut. 46: 263-295.

c. Reproductive Toxicity⁶⁴

Studies identified by U. S. EPA show methylmercury exposure to cause chromosomal aberrations. Both human and animal studies show that methylmercury exposure

causes developmental toxicity. According to U. S. EPA, the most sensitive toxic endpoint in offspring of mothers exposed to methylmercury is neurotoxicity, which can occur in the offspring whether or not any symptoms occurred in the mother during gestation. Manifestations identified by U. S. EPA include:

- Delayed onset of walking
- Delayed onset of talking
- Cerebral palsy
- Altered muscle tone and deep tendon reflexes
- Reduced neurological test scores

***Wildlife

Reproduction is one of the most sensitive toxicological responses, causing effects at very low dietary concentrations (Heinz, 1979; Heinz, 1996; Barr, 1986; Finley, 1978). Concentrations in the egg are typically most predictive of Hg risk to avian reproduction, but concentrations in liver have also been evaluated for predicting reproductive risk. The documented effects of mercury on reproduction range from embryo lethality to sublethal behavioral changes in juveniles at low dietary levels. Effects of Hg include reduced hatchability due to increases in early mortality of embryos; eggshell thinning; reduced clutch size; increased numbers of eggs laid outside the nest, aberrant behavior of juveniles, and potentially may include impaired hearing of juveniles (Fimreite, 1971; Heinz, 1975; Heinz, 1979; Stoewsand, 1971; Scott, 1977).

Barr (1986) indicated reductions in egg laying and territorial fidelity were associated with mean prey Hg concentrations of 0.3 - 0.4 ppm fresh weight; loons established few territories, laid 0-1 egg, and raised no progeny in waters where the mean Hg concentrations of prey exceeded 0.4 ppm fresh weight. The dietary concentrations of MeHg that are required to produce significant reproductive impairment are about 1/5-fold those required to produce overt toxicity in adult birds of the same species (Scheuhammer, 1991). Overall reproductive success in birds can decrease by 35-50% due to dietary MeHg exposure insufficient to cause obvious signs of intoxication in adults. Heinz (1979) fed 0.5 MeHg mg/kg dry wt. (0.1 mg/kg ww) to three generations of mallards. Females laid fewer eggs and produced fewer ducklings. Barr made the same observations in the loon field study mentioned previously where reductions in egg laying and in nest-site and territorial fidelity of the common loon in north western Ontario were associated with maximum mercury residues in eggs of 1.39 mg/kg ww. The loon diet contained from 0.2 to 0.3 mg/kg ww mercury. Heinz also found that ducklings in his multi-generation laboratory feeding study were less responsive to taped maternal warning calls and were hypersensitive to

fright stimulus.

Barr, J. F. 1986. Population dynamics of the Common Loon (*Gavia immer*) associated with mercury-contaminated waters in northwestern Ontario. Canadian Wildlife Service Occasional Paper 56.

Fimreite, N., W. N. Holsworth, J. A. Keith, P. A. Pearce and I. M. Gruchy. 1971. Methyl mercury in fish and fish-eating birds from sites of industrial contamination in Canada. Can. Field Natural. 85: 2211-20.

Finley, M. T. and R. C. Stendall. 1978. Survival and reproductive success of black ducks fed methylmercury. Environ. Poll. 16: 51-64.

Heinz, G. 1975. Effects of methylmercury on approach and avoidance behavior of mallard ducklings. Bull. Environ. Contam. Toxicol. 13: 554-564.

Heinz, G. H. 1979. Methylmercury: Reproductive and behavioral effects on three generations of mallard ducks. J. Wildl. Manage. 43: 394-401.

Heinz, G. H. and L. N. Locke. 1976. Brain lesions in mallard ducklings from parents fed methylmercury. Avian Dis. 20: 9-17.

Scott, M. L. 1977. Effects of PCBs, DDT, and mercury compounds in chickens and Japanese quail. Fed Proc 36: 1888-93.

Stoewsand, G. S., J. L. Anderson, W. H. Gutenmann, C. A. Bache and D. L. Lisk. 1971. Eggshell thinning in Japanese quail fed mercuric chloride. Science 173: 1030-1031.

Scheuhammer, A. M. 1987. Reproductive effects of chronic, low-level dietary metal exposure in birds. Trans. N. A. Wild. Nat. Res. Conf. 52: 658-64.

Scheuhammer, A. M. and D. Bond. 1991. Factors affecting the determination of total mercury in biological samples by continuous-flow cold-vapor atomic absorption spectrophotometry. Biol. Trace Elem. Res. 31: 119-129.

Genotoxicity

Both Hg and MeHg cause chromosome breakage, an effect which is mitigated by H₂SeO₃ (Das et al. 1982, Sayato and Nakamuro 1980). In cultured lung and brain cells from rats, Chinese hamsters and humans, brain cells were more susceptible to MeHg DNA strand

breakage and cytotoxicity than lung cells (Costa et al. 1991, Das et al. 1982, Omata et al. 1986, Sayato and Nakamuro 1980) DeFlora et al reported in an extensive review of the genotoxicity of mercury that Hg compounds often exerted clastogenic effects in eukaryotes, especially by binding SH groups and acting as spindle inhibitors thus causing c-mitosis and resultant aneuploidy and/or polyploidy. Methylmercury compounds were more active than inorganic Hg salts (De Flora, 1994).

Costa, M., N. T. Christie, O. Cantoni, J. T. Zelikoff, X. W. Wang and T. Rossman. 1991. DNA damage by mercury compounds: an overview. In: *Advances in Mercury Toxicology - Rochester Series on Environmental Toxicity*, T. Suzuki, N. Imura and T. W. Clarkson, ed. New York, NY, USA:Plenum Press.

Das, S. K., A. Sharma and G. Talukder. 1982. Effects of mercury on cellular systems in mammals--a review. *Nucleus* 25: 193-230.

Omata, S., H. Kasama, H. Hasegawa, K. Hasegawa, K. Ozaki and H. Sugano. 1986. Species difference between rat and hamster in tissue accumulation of mercury after administration of methylmercury. *Arch. Toxicol.* 59: 249-254.

De Flora, S., C. Bennicelli and M. Bagnasco. 1994. Genotoxicity of mercury compounds. A review. *Mutat Res* 317: 57-79.

Sayato, Y. and K. Nakamuro. 1980. Chromosomal aberration in in cultured human lymphocytes induced by simultaneous treatment with mercury compounds and selenium compounds. *Eisei Kagaku* 26: 99.

3. Reference Exposure Standards

Tables 2-9 and 2-10, respectively, summarize the RfD and MRLs that have been established for methylmercury.

Table 2-9: Reference Doses (RfDs) and Reference Concentrations (RfCs) for Methylmercury⁶⁵

SUBSTANCE NAME	EXPOSURE ROUTE	DURATION OF EXPOSURE	TEST SPECIES	RFC (MG/M ³)	RFD (MG/KG-DAY)
Methylmercury	Oral	Chronic	Human epidemiological studies	Not available at this time.	0.0001

Wildlife reference doses should be included in the table.

Table 2-10: Minimal Risk Levels (MRLs) for Methylmercuric Chloride -- March 1996⁶⁶

SUBSTANCE NAME	EXPOSURE ROUTE	DURATION OF EXPOSURE	TOXIC ENDPOINT	MRL VALUE
Methylmercuric Chloride	Oral	Acute	Developmental	0.00012 mg/kg/day
Methylmercuric Chloride	Oral	Intermediate	Developmental	0.00012 mg/kg/day

Reference doses for wildlife species

Birds (bald eagle, osprey, kingfisher, loon) 21 µg/kg bwt/day

Mammals (mink, river otter) 18 µg/kg bwt/day

Total mercury in unfiltered water : 910 pg/L

Methylmercury in filtered water: 54 pgmeHg/L

USEPA. 1997. Mercury study report to Congress vol 6: An ecological assessment of anthropogenic mercury emissions in the United States. Office of Air Quality Planning and Standards and Office of Research and Development, U.S. Environmental Protection Agency, 4. Bioaccumulation

Contributing factors to methylmercury bioaccumulation are its lipophilic properties, ready absorption in the gastrointestinal tract, and long half-life in the body. "Nearly 100 percent of the mercury that bioaccumulates in fish tissue is methylmercury."⁶⁷

Inorganic and elemental mercury are both toxic, but of the environmentally important forms, methylmercury poses the greatest risk to human health and the environment. This is due both to methylmercury's high toxicity, and the fact that consumption of contaminated fish is the primary route of mercury exposure in humans.⁶⁸ Of the oral routes, methylmercury's poses the greatest risk to humans in non-occupational settings. While metallic mercury has lower reference doses, these are based on the inhalation route of exposure, which is encountered mostly in occupational settings.

Section 2 Key Points:

- Three important forms of mercury exist in the environment: metallic mercury, mercuric mercury, and methylmercury; each has distinct chemical and physical properties, environmental behavior, and toxicology.
- Up to 75 percent of the mercury emitted to the world's atmosphere is of anthropogenic origin, and the world's atmospheric mercury load has increased between two and five-fold since industrialization.
- Mercury is methylated in both the water phase and in sediments.
- Methylmercury bioaccumulates in **both marine and freshwater** food webs, both in the water column and in sediments.
- Inhalation is the most important absorption route for elemental mercury, and neurotoxic effects are its most sensitive toxicological endpoint.
- Mercuric mercury enters the body via inhalation, ingestion, or dermal exposure, and can be methylated by gastrointestinal microbes.
- Methylmercury is a potent developmental and neurological toxin in humans **and wildlife**.
- Methylmercury is well absorbed in the digestive tract.
- Consumption of contaminated fish is the primary route of human methylmercury exposure **in humans**.
- **Consumption of contaminated fish is the primary route of wildlife methylmercury; piscivorous wildlife species, unlike humans, are not protected by fish advisories**

Endnotes

¹ Unless otherwise noted, this data is from:

California Air Resources Board, September 1997. Toxic Air Contaminant Identification List Summaries. p. 626.

² U.S. Environmental Protection Agency, Office of Science and Technology and Office of Water, January 2001. Water Quality Criterion for the Protection of Human Health: Methylmercury—Final. Publication EPA-823-R-01-001, pp. 1-2 - 1-3. Internet web site, accessed June 10, 2001:
<http://www.epa.gov/waterscience/criteria/methylmercury/criteria.html>

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Section 5: Waste Contribution to the Mercury Environmental Burden

I. Introduction

This section focuses on the contributions of waste to the mercury emissions into the air, water and land. Information and data regarding waste-derived sources and their quantities into the air, water and land are presented in the first subsection, Mercury Anthropogenic Sources and Emissions. It is followed by an assessment of those mercury emissions in the following subsection, Mercury Environmental Burden Assessment. Waste combustion sources are emitted in significant quantities relative to California waste-derived sources. Identified water waste-derived mercury sources include legacy wastes, dentistry, and fluorescent lights. Land sources include disposal of mercury-containing products. A qualitative assessment of the quantities of waste-derived sources of mercury into the environment was done and it was estimated that

- 1.3 short tons of mercury from lamps would potentially be disposed in 2001.
- 2.24 short tons of mercury from waste-derived sources were emitted into the atmosphere in 2000.
- 0.4 short tons of mercury in auto shredder fluff were disposed in landfills in 2001.
- 118 pounds of mercury from dental offices exited the POTWs in 2000.
- 2.2 short tons of dental mercury were recycled or (land) disposed in 2000.

Although California agencies are working to reduce or control mercury emissions into the environment, mercury's mobility has continued to be an environmental issue, as evidenced by fish consumption advisories. DTSC is considering additional steps to control mercury emissions to land.

II. Mercury Anthropogenic Sources and Emissions

The following subsection focuses on the mercury contained in wastes, trends in waste mercury content, and the relative contribution of disposal of this waste to the total environmental mercury burden. Since the beginning of the industrial age, an estimated three-fold increase in the global environmental mercury burden has been attributed to human activities.¹ Mercury is mobile within and between air, water, and soils and is a public health and environmental concern. It follows that any steps that limit or control the amount of anthropogenic mercury entering the environment will yield benefits. This includes controlling the amount of mercury used as a raw material for industrial processes and consumer products through pollution prevention techniques, such as source reduction or substitution, or through indirect means, such as banning the sales of mercury-containing products, or imposing disposal restrictions of mercury-containing waste.

A. Anthropogenic Sources - Raw Material

1. Domestic Supply Trends²

An overall review of the supply of mercury is important in understanding the trends of its production and resulting release to the environment. In the USGS 2000 study of the materials flow of mercury from 1970 to 1998, Sznopek and Goonan identify "three

different time periods, each characterized by different market dynamics” were identified. The first of these periods lasted from 1970 to 1986. During this time, “. . .U. S., primary mercury mine production and net imports contributed significant amounts to the mercury market”.

During the second period, which began in 1986 and lasted until 1992, the United States apparent mercury supply saw a rapid decrease, due in large part to the adoption of legislation to eliminate mercury in batteries. Battery manufacture accounted for 54 percent of the demand for mercury in 1984, but for only 2 percent of the mercury demand in 1992. During the same period, mercury was eliminated as a fungicide in paints. Fungicide use accounted for 16 percent of the demand from mercury in 1989; by 1992, it's accounted for none of the nation's demand. Apparently due to the dramatic drop in demand for mercury, the United States actually reversed the trend of large imports of mercury to become a net exporter of mercury beginning in 1989 and lasting through 1994. Mine production of primary mercury in the United States ceased in 1991.

The third period identified in the USGS 2000 Study lasted from 1993 to 1998. It was characterized “. . . by increases to consumer and producer stocks, increasing net imports, no primary mine production, and greatly expanded secondary mercury production, supported by . . . legislation mandating mercury recycling”.³

2. Domestic Consumption (Demand) Trends

Figures 5-1 and 5-2 are reproduced from the USGS 2000 Study. Weights are reported in metric tons in the two figures, but in the text of this report, all weights were converted to short tons for discussion purposes.* Figures 5-1 and 5-2 show the corresponding drop in mercury consumption during the late 1980s until the early 1990s.

* One metric ton equals 1000 kilograms, or 2200 lbs. One short ton (2000 lbs.) equals 0.907 metric tons.

Figure 5-1: U. S. Industrial Reported Consumption of Mercury (1970-1997)

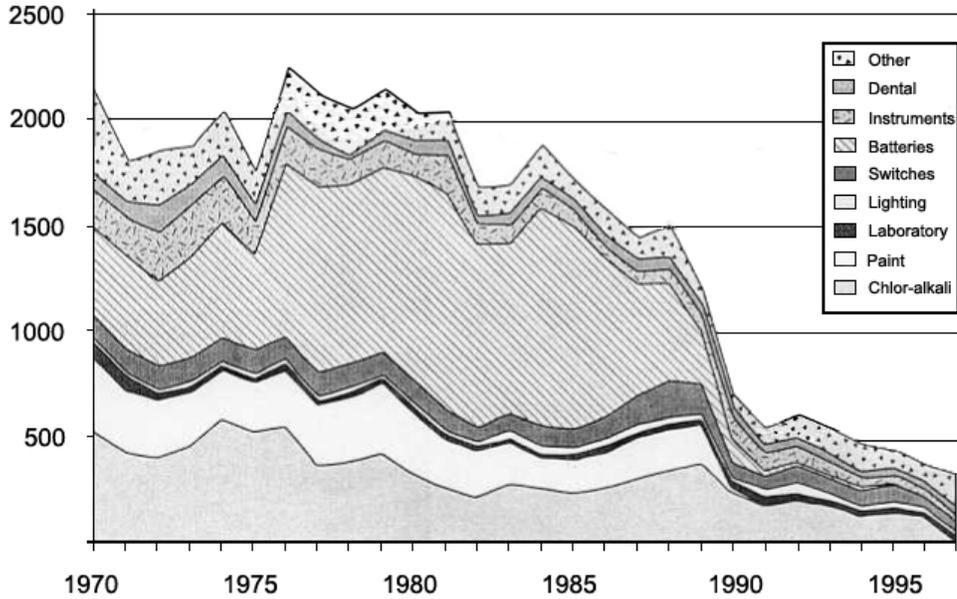


Figure 5-1 shows a steep drop of mercury consumption from the late 1980s through the early 1990s. This trend has continued, although the sharp downward slope has eased. The decrease in demand has been significant in most categories, except for dental, switches, lighting, and laboratory uses.

Figure 5-2: U. S. Apparent Supply and Reported Consumption of Mercury (1970-1998)

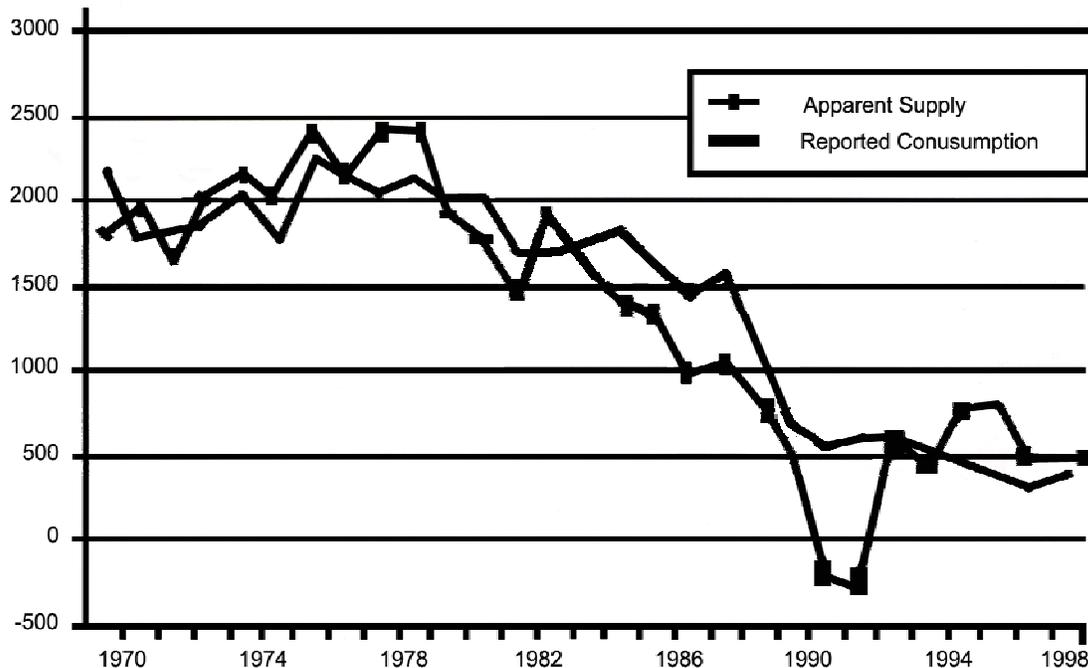


Figure 5-2 shows the corresponding supply and demand graphs for a similar period. The trend depicted in this graph supports the trend shown in Figure 5-1. The two figures show the United States consumed approximately 2200 short tons of mercury per year during the period from 1970 through 1986, then dramatically reduced its consumption to approximately 550 short tons per year between 1986 and 1992. The apparent supply closely follows the mercury consumption, except for the period during the early 1990s, when the United States was a net exporter of mercury.

The U.S. EPA 1997 Study has estimated domestic mercury consumption in 1989, during the second period identified in the USGS 2000 Study, to be 1336 short tons.⁴ The U.S. EPA 1997 Study's estimate is in close agreement with the USGS 2000 Study's estimate for 1990: 1,354 short tons.⁵

Figures 5-1 and 5-2 show that domestic mercury consumption dropped from more than 2426 short tons in 1976 to less than 441 short tons in 1998.⁶ As the use of mercury continues to decline, mercury releases to the environment incidental to the manufacture, use, and disposal of products can also be expected to fall. Recent developments are likely to increase the downward trend in mercury consumption. For example, legislation introduced in 2001 is pending in many states that would effectively restrict the manufacture by prohibiting the sale of a certain mercury-added products (refer to Appendix A: Summary of Nationwide Mercury Efforts). The use of mercury in other products, including pesticides, mildewicides for paints, and many batteries, has already been eliminated.

3. Mercury Flow Trends⁷

According to the USGS 2000 Study, primary mine production of mercury fell from 494 short tons in 1990 to zero in 1996.⁸ During the same period, secondary production of mercury increased to 492 short tons, more than four times the level in 1990. In 1990, the United States government sold 270 short tons of mercury from its stockpiles. United States government mercury sales were suspended in 1994 and have apparently not resumed. It appears that secondary mercury production has replaced primary mercury production.

According to the USGS 2000 Study, the total mercury flows to industry fell significantly. They were reduced from 784 short tons in 1990 to 410 short tons in 1996, as shown in Figure 5-2.⁹ Figure 5-1 shows a decrease in the mercury flows to the following industrial sectors:

- dental (30 percent),
- laboratory (38 percent),
- measurement and control devices (62 percent),
- wiring and switches (30 percent),
- lighting (66 percent),
- batteries (100 percent), and
- chlor-alkali plants (45 percent).

The most dramatic decrease was mercury use in batteries, which went from 116 short tons in 1990 to virtually none in 1996.

B. Air Emissions

Fossil fuel combustion emitted 84 short tons of mercury to the nation's air in 1996. Of this total, 73 tons were caused by the combustion of coal.¹⁰ Oil and gas combustion for residential and non-residential space heating emitted 11 short tons into the air, while waste incinerators emitted 60 short tons.¹¹ The three main types of waste incinerators were: municipal waste combustors which emitted 30 short tons, medical waste incinerators which emitted 17 short tons, and hazardous waste combustors and cement kilns which emitted 12 short tons.¹² Table 5-1 displays these emissions.

Table 5-1: U.S. Mercury Emissions from Combustion Sources, 1996¹³

Source	Mercury Emissions (Short Tons)
Coal burning	73
Oil/gas combustion	11
Municipal waste combustion	30
Medical waste combustion	17
Hazardous waste combustion	12

In its 1997 report to Congress, U.S. EPA reported estimated United States mercury air emission rates for a number of source categories. Although they warn that their numbers are intended to be only estimates, U.S. EPA believes that “they ... provide insights into the relative magnitude of emissions” from the different sources.¹⁴ In each of the years 1994 and 1995, U.S. EPA reports that United States atmospheric mercury emissions totaled 158 short tons.¹⁵ Of this total, “approximately 87 percent is from combustion point sources, 10 percent is from manufacturing point sources, 2 percent from area sources, and 1 percent is from miscellaneous sources”.¹⁶ Of the non-combustion sources, the largest national contributor was the chlor-alkali industry, which emitted 4.5% in 1994-1995. For the purpose of the U.S. EPA inventory, the nonhazardous waste incinerating Portland cement operations (3.1%) were counted as a manufacturing process. Pulp and paper manufacturing contributed 1.2%.¹⁷

Air releases from waste incineration decreased from 110 short tons in 1990 to 58 short tons in 1996. This was apparently due to a reduction in the amount of mercury contained in products as well as an increase in the efficiency of air emission controls.

Data collected by ARB and presented in Section 3 are summarized in Table 5-2 reflect those, which would include traditional waste-derived sources totaling to 4490 pounds/year or 2.24 short tons/year.

Table 5-2: California Waste-Derived Air Emissions for 2000

Waste-Derived Source	Mercury Emissions pounds/year
Industrial Processes (cement manufacturing)	2500
Agricultural and Rangeland Prescribed Burning (waste burning)	440
Fluorescent Tube Breakage	370
Electric Utilities (municipal waste fueled cogeneration plants)	900
Other (waste disposal, landfills, soil remediation, sewage treatment, medical and municipal waste incinerators)	280
	4490 pounds/year or 2.24 short tons/year

Although a direct comparison to national data cannot be done due to differences in sources and the differences in reporting requirements, a rough comparison was made with national waste combustion sources with California waste-derived sources. Nationally, waste combustion sources contributed to 59 short tons, while in California, the waste-derived sources contributed to 2.24 short tons. An estimate based on a 12% per capita the national combustion sources would have yielded an estimate of 7 tons of air emissions, while the California air emissions for 2000 yielded 2.24 tons, significantly

less. This difference may be attributed to the different years in which the national estimates were collected and compared. However, it is more likely that California has less medical and municipal waste incinerators, and no offsite hazardous waste incinerators. There are three onsite boiler/industrial furnaces that are permitted to burn hazardous waste, but one has not operated since it has been permitted by DTSC.

Mercury emission sources that were reported by the ARB in other source categories that emit more than 100 pounds of mercury per year include: geothermal sources, petroleum-related manufacturing, general manufacturing, fuel combustion sources, off-road and on-road mobile sources.¹⁸

C. Water Emission (Sources)

A large proportion of California's aquatic mercury burden originates from legacy waste from inoperative mercury and gold mines. As it is slowly mobilized from sediments, this 'legacy' mercury is carried from parent water bodies to the other water bodies into which they drain. Other sources of mercury into water bodies are atmospheric deposition, remobilization of historically polluted sediments through erosion, and wastewater discharges from point source discharges.¹⁹ The mercury contained in waste can make its way into California's waters by leaching and runoff from landfills, by atmospheric deposition, and via the sewer system.

It is suspected that in urbanized areas, dental amalgam may be a major contributor of mercury to wastewater that is treated by the POTWs. In a study conducted by the city of Palo Alto, it was found that in 2000, approximately 24 pounds of mercury entered the wastewater treatment plant, with about 20 pounds originating from dental amalgam (dental offices and human wastes).²⁰ Based on the information contained in the *Mercury Headworks Analysis for 2000* (Palo Alto Mercury Headworks 2000 Analysis) that was prepared for the Palo Alto Regional Water Quality Control Plant (RWQCP), 11.4 pounds per year enter the POTW for treatment from 170 dentists in the Palo Alto RWQCP service area.²¹ Using this data as a basis for determining the impact in California for the 20,000 active dentists in California and that 12% of the dentists do not use amalgam, an estimated 1,180 pounds of dental amalgam enters California's POTWs for treatment. POTWs mercury removal efficiency typically is 90%, resulting in discharges to water sources of 118 pounds in California.²²

Abu-Saba, et al., in their *Watershed Management of Mercury in the San Francisco Bay Estuary: Total Maximum Daily Load Report to U.S. EPA*, June, 2000 (San Francisco Bay TMDL 2000 Report), has estimated that breakage of fluorescent light bulbs in landfills in their locale may contribute from 22 to 286 pounds per year as air emissions and deposits mercury into the San Francisco Bay.²³

D. Land Emissions (Disposal)

The USGS 2000 Study states that the amount of mercury disposed in landfills (excluding soil amendments) dropped from 832 short tons in 1990 to 325 short tons in 1996.²⁴ The U.S. EPA 1992 Study's estimate of landfill disposal of mercury in 1989 is in fair agreement with this figure. The U.S. EPA 1992 Study reported that in 1989, 709

short tons of mercury were discarded in municipal solid waste in the United States.²⁵ Summaries of the amount of mercury disposed are shown in Tables 5-3 and 5-4 below. The tables are reproduced from the U.S. EPA 1992 Study.²⁶

Tables 5-3 and 5-4 show U.S. EPA's projections of mercury discards for 2000, based on data collected from 1970 to 1989. Table 5-3 lists the contributions to mercury in the municipal solid waste (MSW) nationwide from each of the largest mercury-containing product categories; Tables 5-4 lists the relative contributions of each of these categories. The amount of mercury discarded in California for 2000, and the relative contributions of the various product categories in are estimated in Tables 5-3A and 5-4A; these tables are adjacent to Tables 5-3 and 5-4, respectively. As in Section 1, the calculated values in Tables 5-3A and 5-4A are based on the assumption that California's discards are representative of the nation's discards, and that the State's population represents 12 percent of the United States population.

This table is confusing, because the units in Table 5-3 are in short tons, but Table 5-3A says "per capita" but no units are given.

Table 5-3: DISCARDS¹ OF MERCURY IN PRODUCTS IN THE MUNICIPAL SOLID WASTE STREAM 1970 TO 2000 (In short tons ²)²⁷

Products	1970	1980	1989	2000 (Projected)
Household Batteries	310.8	429.5	621.2	98.5
Electric Lighting	19.1	24.3	26.7	40.9
Paint Residues	30.2	26.7	18.2	0.5
Fever Thermometers	12.2	25.7	16.3	16.8
Thermostats	5.3	7.0	11.2	10.3
Pigments	32.3	23.0	10.0	1.5
Dental Uses	9.3	7.1	4.0	2.3
Special Paper Coating	0.1	1.2	1.0	0.0
Mercury Light Switches	0.4	0.4	0.4	1.9
Film Pack Batteries	2.1	2.6	0.0	0.0
TOTAL DISCARDS	421.8	547.5	709.0	172.7

Table 5-3A

California 2000 (Per Capita Projection)*
11.8
4.9
0.1
2.0
1.2
0.2
0.3
0.0
0.2
0.0
20.7

¹ Discards before recovery.

² 1 Short Ton equals 2000 pounds

Source: Franklin Associates, Ltd.

* Based on assumption that California's population is 12% of the national population

As shown in Table 5-3, U.S. EPA estimated that, in 1989, 709 short tons of mercury were discarded to municipal solid waste.²⁸ Batteries accounted for 87.6 percent (621.2 short tons) of this total and lighting accounted for 3.8 percent (26.7 short tons), as shown in Tables 5-3 and 5-4.²⁹

Table 5-4: DISCARDS¹ OF MERCURY IN PRODUCTS IN THE MUNICIPAL SOLID WASTE STREAM 1970 TO 2000 (In Percent of Total Discards)³⁰

	1970	1980	1989	2000 (Projected)
Household Batteries	73.7	78.4	87.6	57.0
Electric Lighting	4.5	4.4	3.8	23.7
Paint Residues	7.2	4.9	2.6	0.3
Fever Thermometers	2.9	4.7	2.3	9.7
Thermostats	1.3	1.3	1.6	6.0
Pigments	7.7	4.2	1.4	0.9
Dental Uses	2.2	1.3	0.6	1.3
Special Paper Coating	0.0	0.2	0.1	0.0
Mercury Light Switches	0.1	0.1	0.1	1.1
Film Pack Batteries	0.5	0.5	0.0	0.0
TOTAL DISCARDS	100.0	100.0	100.0	100.0

Table 5-4A

California 2000 (per Capita Projection)*†
6.8
2.8
0.0
1.2
0.7
0.1
0.2
0.0
0.1
0.0
12.0

¹ Discards before recovery.

Source: Franklin Associates, Ltd.

*Assumption based on California's population is 12 % of the nation's population

Tables 5-4 shows that U.S. EPA projected changes in the relative contribution of batteries and lamps, the two largest categories of mercury-containing products, to the total amount of mercury in discarded products. U.S. EPA projected that the contribution of batteries to the total amount of mercury in MSW would significantly decrease: from 87.6% in 1989 to 57% in 2000.³¹ U.S. EPA also projected that the contribution of the disposal of electric lighting would increase from 3.8% to 23.7% during the same period.³² Taken together, batteries and electric lighting were projected to account for 80.7% of the mercury in discarded products in 2000. Based on per capita projections for batteries and electric lighting, California would be expected to have 9.7% of the nation's battery and electric lighting discards in 2000, which represents 16 short tons of mercury into California's landfills (See Table 5-3A).

In U.S. EPA's summary of mercury in discarded products, the contribution from fever thermometers and thermostats did not show signs of decreasing between 1970 and 1989, nor did U.S. EPA project significant reductions by 2000. The amount of mercury in discarded fever thermometers was 16.3 short tons in 1989 and was projected to be 16.8 tons in 2000.³³ The amount of mercury discarded in thermostats was 11.2 and

† California's contribution to the national mercury discharge.

10.3 short tons for the respective years³⁴ (see Table 5-3). Based on the previous assumptions, California would be projected to discard 3.2 short tons of mercury from fever thermometers and thermostats in 2000, representing 1.9% of the nation's total mercury discards (see Tables 5-3A and 5-4A).

The amount of mercury discarded nationally from light switches showed no change between 1970 and 1989, but was projected to increase to 1.9 short tons (1.1%) in 2000.³⁵ Similarly, the California estimate in 2000 would be 0.2 short tons entering California's waste stream and 0.1% of the nation's total mercury discards.

Mercury discards in MSW peaked in 1986 and are declining.³⁶ The U.S. EPA analysis agrees with the USGS 2000 Study's analysis in that a significant decrease was expected from batteries and paints. The U.S. EPA identified electric lighting and mercury light switches as the only mercury products with increasing quantities.³⁷ Taken together, the disposal of these two product categories was predicted to have contributed 24.8 percent of the total mercury discarded nationwide in 2000.

E. Fluorescent Lamp Data

The USGS 2000 Study reported that mercury content in fluorescent lamps shows a linear decreasing trend.³⁸ In 1990, the reported content was 46 milligrams per lamp, followed by 38 milligrams in 1991, 34 milligrams in 1992, 30 milligrams in 1993, 27 milligrams in 1994, and 23 milligrams in 1995.³⁹ The projected figure for 1996 was 19 milligrams per lamp.⁴⁰ U.S. EPA reported that the average fluorescent lamp had 75 milligrams of mercury from 1970 through 1984, as compared to 55 milligrams for lamps manufactured after 1985.⁴¹ This data confirms the linear decrease in average mercury content from 1985 through 1995 that is seen in the USGS 2000 Study's data for the same time period.

Data cited by the USGS 2000 Study show a 35 percent decrease in mercury content in fluorescent lamps between 1985 and 1995.⁴² However, calculations based on data from the USGS 2000 Study and the U.S. EPA 1992 Study show a much steeper drop: a reduction from 55 mg per tube in 1985 to 23 mg per tube in 1995, representing a 58% decrease. The San Francisco Bay TMDL 2000 Report cites data that commonly-used T8 fluorescent tubes contain approximately 10 mg of mercury each, while larger-diameter T12 tubes contain 21 mg per bulb, on average.⁴³ This indicates that the rate of the reduction in the mercury content of lamps may have slowed; technology may have reduced the mercury content of lamps to the point that further reductions would adversely affect lamp performance.

There is no discussion of compact fluorescent lamps here. CFLs contain less mercury than conventional fluorescent lamps of comparable light output; last longer, therefore requiring less frequent replacement; and use less energy, thus decreasing power plant output, and its corresponding mercury air emission. It would be worth calculating the decreased contribution to the mercury waste stream that could be achieved with an increase the use of CFLs.

U.S. EPA estimates that 26.7 tons of mercury was disposed in electric lighting in 1989.⁴⁴ Assuming that California's lamp usage and disposal patterns are proportional to national usage and disposal, and considering that California's population is 12% of the national population, it is estimated that California discarded 12% of the 26 tons, or 3.1 tons of mercury to MSWs from lamps in 1989. Based on information provided by the National Electrical Manufacturer's Association, the approximate amount of mercury originating from fluorescent lamps that will impact California in 2001 will be 2686 pounds or about 1.34 short tons.⁴⁵ This is about 45% less than the 2000 estimate of 4.9 short tons projected in Table 5-3A. The 2001 estimate is based on the number of lamps sold in 1996 with an estimated 16 milligrams of mercury and based on a five-year life expectancy.

F. Dentistry

The use of mercury in dental amalgams is being seriously debated worldwide. Governments that have taken steps towards eliminating or limiting amalgam use include Sweden, Germany, Denmark, Norway, Finland, Canada, and Austria.^{46,47,48} In California, Senate Bill 134 (Chapter 532, Statutes of 2002) requires a disclosure form signed by all patients regarding the comparative risks and efficacy of various types of dental restorative materials. Congresswoman Watson introduced HR 413 in April 2002 to require the same type of disclosure on a national level. However, national data in Table 5-3 show that mercury discards to MSW from dental uses are declining. In 1989, 4.0 short tons were disposed; U.S. EPA projected that 2.3 short tons would be discarded to MSW in 2000.⁴⁹ Using these figures to project the same data in California, California dentists would have contributed 0.48 short tons (960 pounds) in to MSW in 1989 and estimated 0.28 short tons (560 pounds) in 2000. California's dental amalgam waste is projected to have contributed 0.2 percent of the nation's total mercury discards in 2000.

Based on information contained in the *Mercury Headworks Analysis for 2000* that was prepared for the Palo Alto RWQCP, an average of 0.45 grams per day of dental amalgam scrap is captured by dental offices in chairside traps, vacuum screens, or other capture method.⁵⁰ Using again that there are 20,000 active dentists in California and 12% of the dentists do not use amalgam, there were 2.2 short tons of dental mercury that was disposed or recycled in California in 2000. The California quantity is based on dental mercury generated rather than land disposed and although not directly comparable, this quantity is greater than the projected estimate for 2000 that would have been disposed to California landfills.

G. Data Limitations

The air and land emissions reported in 1996 from the USGS 2000 Study's data and the U.S. EPA 1992 Study's data are applicable to the United States as a whole. The U.S. EPA 1992 Study cautioned that the data should not be construed to be representative of mercury in MSW in a particular locality, as there are variations in waste composition and waste management practices.⁵¹ The report also cautioned that the estimates are often based on assumptions. The U.S. EPA 1992 Study also excluded a number of

nonhazardous wastes (municipal sludges, oil and gas production wastes, and mining wastes, for example) from their calculations.

The U.S. EPA 1997 Study report acknowledged that there are “considerable uncertainties regarding the levels of natural and re-emitted mercury emissions.”⁵² This makes “an assessment of the relative public health and environmental impact that can be attributed to current anthropogenic emissions... (very) complicated...”⁵³ U.S. EPA’s external review panel estimated that the missing sources from its report could contribute as much as 20 percent more mercury emissions to the United States total.⁵⁴ However, the U.S. EPA 1997 Study’s estimate compares favorably (within 10%) with two other studies done for 1990, and the 1994-1995 national baseline study

Similarly, some of the California estimated projections will have uncertainties as they were calculated based on 12% of national data, a per capita basis, and the assumption that California’s consumption and discards is ~~en~~ representative of the nation. When California specific data were available, these were included for assessment purposes.

III. Mercury Environmental Burden Assessment

The data presented above indicate that mercury’s use as a raw material is declining, as shown by the decreases in supply and demand of mercury. This is attributed to declining mercury uses in industry and products resulting from regulatory efforts to limit or decrease mercury uses. Secondary production (recycling) has completely supplanted primary production of mercury from ore, and appears to be adequate to meet the reducing demand for the metal. There are, however, existing stockpiles of mercury as a raw material that may become a long term storage or disposal issue when the supply greatly surpasses the demand for mercury.

Nevertheless, it follows that if there is a declining usage of mercury in industry and products containing mercury, there will be a downward trend in the amount of mercury-containing waste entering the land from direct disposal. Additionally, as future regulatory efforts to control and decrease emissions to air (air pollution control devices), water (Clean Water Act and TMDL efforts), and land disposal (hazardous waste treatment before land disposal) continues, the mercury industry and consumers will be considering the cost effectiveness of the continued use of mercury.

While the use of mercury has continued to drop, it is clear that the environmental mercury burden remains unacceptably high. Past activities have mobilized mercury in the environment, where it persists and continues to pose risks to public health and the environment. This fact is evidenced by numerous sport fish consumption advisories issued in California and in other states, by the mercury-contaminated sites that require mitigation, and by the numerous legislative and regulatory efforts to reduce the amount of mercury that enters the environment **throughout** the nation and in California (see Appendix A: Summary of Nationwide Mercury Efforts).

A. Air and Water Waste Burden Assessment

Air emissions from anthropogenic sources are decreasing, due not only to decreases in industrial uses, but due to increased efficiency of air pollution control devices. The latter factor has been driven by statutes and regulations, such as the California's Air Toxic "Hot Spots" program that are intended to reduce air pollution with toxic substances. Nationally, the mercury contribution from waste combustors (municipal, medical, and hazardous waste combustors) to air emissions in 1996 was 60 short tons while in California, the 2000 mercury waste-derived source emissions were 2.24 short tons⁵⁵.

Controlling mercury entering water sources continues to pose a challenge as indicated by efforts in the San Francisco Bay TMDL 2000 Report and the Palo Alto Mercury Headworks 2000 Analysis. Point source wastewater discharges from industry and POTW, although controlled, are suspected to contribute to the mercury deposition in the Bay and cause impairment to the waters and water sediments, which ultimately result in mercury fish consumption advisories. Other statewide efforts to address mercury in the water bodies are in the initial stages (for example, Central Valley Regional Water Quality Control Board's TMDL for Clear Lake).

The San Francisco Bay TMDL 2000 Report notes efforts to estimate the amount of mercury from lights from breakage at the landfill, which may contribute to the bay's mercury loading through atmospheric deposition. The report suggests that partnerships with manufacturers to further reduce mercury in lighting or efforts to ensure 100 percent recycling instead of landfill disposal as two possible mechanisms to reduce atmospheric mercury emissions. Another suspected source of mercury in the San Francisco Bay is dental amalgam waste. Mercury has been found in POTW effluents, in spite of the fact that the influent waste is extensively treated prior to discharge, attaining effluents with mercury concentrations from 5-7 ng/L in advanced treatment plants to 15-25 ng/L in secondary treatment plants. While mercury removal is efficient, a better strategy is to reduce the potential 1180 pounds of mercury influent as much as possible with mercury alternatives as discussed in Section 4 or pollution prevention techniques, such as additional mercury traps. The resulting mercury reduction entering the POTWs will reduce the effluent after treatment.

Another major source of mercury contamination noted in the San Francisco Bay TMDL 2000 Report is legacy waste from past mercury mining. The report states that, in order to achieve the proposed TMDL goals, all efforts to reduce introduction of mercury in the bay will be needed, including increased current efforts.

California's waters are under the regulatory authority of the California State and Regional Water Quality Boards. Efforts to control the discharges into sewers and POTWs are a joint effort of the State and Regional Water Quality Control Boards, DTSC and their delegated local implementing agencies.

For instance, as noted in San Francisco Bay TMDL 2000 Report and the Palo Alto Mercury Headworks Analysis, amalgam and fluorescent lights are considered sources of mercury in the Bay and in wastewater. DTSC oversees the management and

disposal of amalgam waste and most mercury-containing fluorescent tubes. Amalgam waste from dental offices is considered hazardous waste and most dental offices recycle the waste amalgam under the scrap metal exemption. However, it has recently come to DTSC attention that during the processes that generate the amalgam waste during dental operations, small amounts enter the POTW system from each dental office, totally at an estimated at 1180 pounds of mercury from California dentists. Each dental office may contribute insignificant amounts of amalgam into the POTW, but the amount of dental offices in the area may add up to a significant amount of mercury entering the POTW. As noted in the Palo Alto Headworks Analysis, about 80% of mercury entering wastewater treatment originates from dental amalgam sources.⁵⁶

In like fashion, most fluorescent tubes currently contain mercury in concentrations that are considered hazardous waste and must be managed accordingly. However, as manufacturing industry progresses and the mercury concentration in lighting is reduced to the point that the lighting waste is below the hazardous waste threshold, the consequences may equate to a significant source of mercury to air, water and land. That is, the quantity of lighting waste, along with their reduced concentrations of mercury to nonhazardous waste levels, may add up to a significant amount of mercury, adding to the total mercury burden in air and water, as well as to their impact to direct land contamination, which is discussed below.

B. Land Burden Assessment

Since the mid 1980s, appropriate land disposal of mercury-containing waste has been determined by an assessment of the hazardous waste identification criteria, whether a federal "listed" hazardous waste, or a mercury characteristic waste by the TCLP, WET, or TTLC. If the mercury in the waste is determined to be a hazardous waste, the land disposal is controlled, as well as its storage, transportation, treatment, and recycling. The oversight of this regulatory scheme falls within DTSC.

However, not all waste falls within this regulatory scheme and under DTSC. For instance, a waste may meet hazardous waste criteria, but be exempt from regulation by DTSC because of a statutory or regulatory exemption.

In evaluating the wastes that are under the authority of DTSC as discussed in the Land Emissions (Disposal) of this section, many of the wastes meet current hazardous waste identification criteria and must be managed in accordance to requirements for hazardous waste. This includes the estimated projection of 20 short tons of mercury. These include switches, batteries, and thermometers, paints and most mercury-containing electric lighting. The mercury discards in Table 5-3 and 5-4 are managed as hazardous wastes in California and should not be entering Class III landfills. As a general statement, most consumer product wastes with little or light housing may be a hazardous waste since the mercury concentration in the product would be distributed over the total weight of the waste. For instance, mercury in paints would be considered a hazardous waste, but if the mercury-paint was on wood debris, the concentration of mercury may not be sufficient in relation to the total wood waste to be considered a hazardous waste for controlled management and disposal.

Wastes that may be nonhazardous or **are** expected to be nonhazardous are those wastes that exist in large or heavy housing or in equipment where the mercury cannot be removed or is difficult to remove. Examples of these types of wastes are measuring equipment, such as manometers or barometers which are made with heavy and/or large housing and where the mercury measuring device is **so** securely housed that dismantling is difficult; toys, games, novelty items with embedded mercury batteries or switches; and cars containing mercury switches. Because the California hazardous waste criteria is based on WET-soluble and total concentrations, the mercury is "diluted" with the housing and may be determined to be nonhazardous for disposal.

In California law, appliances are diverted from disposal in Class III municipal landfills and are recycled for their scrap metal. This law also requires that mercury switches/devices be removed before recycling the metal. Currently, the law does not apply to automobiles, which are also recycled for their metal. Consequently, non-ferrous waste generated from shredding automobiles is contaminated with mercury, but is "diluted" to nonhazardous waste concentrations due to the large mass of each automobile. If mercury switches were removed before shredding automobiles and properly managed as a hazardous waste, a significant amount of mercury could be diverted from Class III landfills.

DTSC's Auto Shredder Initiative has estimated that 700,000 automobiles are shredded in California each year. Each car has two mercury switches, containing an average of 500 and 1000 mg of mercury each.⁵⁷ Assuming that none of these switches are currently removed prior to disposal, and the amount of mercury disposed to non-hazardous waste landfills via auto shredder waste, a mixture of appliance and automobile shredder waste, is between approximately 0.75 and 1.5 short tons. The DTSC Auto Shredder Initiative sampling effort has shown that there **are** 300,000 tons of auto shredder waste with a total of 0.93 short tons of mercury. Of the 0.93 tons of mercury, it is estimated that 0.4 short tons originated from automobiles (47% of the shredding operation are from automobiles) with an undetermined amount being emitted to the air during storage or during the shredding operation.

Information from "nonhazardous" fluorescent lamps is limited. It is estimated in 2001 that California will have a disposed of potentially 1.34 short tons of mercury from all fluorescent lamps.⁵⁸ DTSC has received anecdotal information indicating that 25% of the mercury lamps disposed in California are "nonhazardous" fluorescent lamps; however, confirmation of this information is needed.

Suspected "nonhazardous" waste, such as, toys, games, novelty items, nonhazardous electrical lighting waste, measuring equipment, and painted debris, etc., enter a Class III municipal landfill. Nonhazardous waste treatment, storage, transportation and disposal requirements are not the same as hazardous waste requirements. This may cause potential for mismanagement occurrences during their handling, storage, transportation, and disposal, which may result in potential breakages, spills, and leaks to the land and air. Small quantities of mercury spills and leaks during handling and storage may cause

direct land contamination over time. This may result in a contaminated site, which may require clean up to protect public health. Mercury air emissions due to breakage, spills, and leaks are uncontrolled and cause an incremental increase in the inhalation hazard. Mercury may enter the water due to breakage, spills, leaks and improper storage or disposal and enter storm drains and ultimately the open waters. The quantities of these wastes are unknown at this time; however, there has been an incident involving a contractor lighting change out and dumpster disposal, which resulted in many fluorescent lights broken near a storm drain.⁵⁹

Clearly, as much as California has controlled mercury releases to air, water, and land, to protect public health and environment, the mercury burden and its mobility to travel between environmental media, is still an environmental issue as evidenced in water pollution and fish consumption advisories. Additional controls are necessary to protect public health and environment. Currently, it is easier to dispose of mercury-containing waste rather than recycling the waste and there is no incentive to recycle. Water agencies are considering additional measures to protect California's water from mercury sources in their TMDL effort. California legislation in 2001 has been introduced to ban sales of mercury-containing products in California as well as "encourage" the removal of mercury light switches in automobiles. Nationwide and state mercury organizations exist to address mercury in the environment.

California agencies overlap and affect each other's primary responsibility in protecting public health and environment in regards to mercury in our environment. Each agency is charged to protect public health and environment to the extent their regulatory authority allows them. The California Environmental Protection Agency has charged these agencies to work in cooperation with each other, to address public health and environmental issues. As such, to provide additional safeguards, encourage pollution prevention and promote recycling, DTSC is recommending the regulatory concept to identify intentionally added mercury-containing products as a hazardous waste when they are discarded.

Section 5 Key Points:

- An estimated three-fold increase in the global environment mercury burden has been attributed to human activities.
- From 1970 to 1986 U. S. conducted mercury mine production and imported mercury.
- From 1986 to 1992 mercury supply and use is decreased and the United States exported mercury.
- From 1993 to 1998, the United States does no primary mercury mine production and uses secondary production of mercury to meet its supply needs.
- Domestic mercury consumption dropped from more than 2426 short tons in 1976 to less than 441 short tons in 1998.
- Fossil fuel combustion emitted 84 short tons of mercury to the nation's air in 1996, with waste incinerators emitted 60 short tons.
- California's air emissions from waste-derived sources are 2.24 tons in 2000.

- The ARB estimates that 450 pounds of mercury air emissions were derived from broken fluorescent tubes.
 - In 1994 and 1995, approximately 87 percent of the nation's atmospheric mercury emissions were from combustion point sources.
 - A large proportion of California's aquatic mercury load originates from legacy waste from inoperative mines.
 - An estimated 22 to 286 pound per year from fluorescent lights potentially enters the San Francisco Bay.
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- The USGS estimated that the amount of mercury disposed in landfills dropped from 832 short tons in 1990 to 325 short tons in 1996.
 - Household batteries and lighting comprise the majority of the discards of mercury in products in the municipal solid waste stream from 1970 to 2000.
 - U.S. EPA's study showed that the mercury contribution from fever thermometers and thermostats did not show signs of decreasing between 1970 and 1989. No significant reductions were projected by 2000.
 - The mercury content in fluorescent lamps has decreased significantly since 1985 to 1995 and is slowly decreasing, indicating that further decreases in mercury will affect lamp life.
 - U.S. EPA estimates that 26.7 tons of mercury was disposed in electric lights in 1989.
 - California estimates that 1.3 short tons of mercury from fluorescent lamps will be disposed in 2001.
 - California dentists generated an estimated 2.2 tons of mercury from dental amalgam that was disposed or recycled and 118 pounds of mercury from dental offices exited the POTWs into waterways.
 - While the use of mercury has continued to drop, the environmental mercury load remains unacceptably high. This is evidenced by numerous sport fish advisories, by the mercury-contaminated sites, and by the numerous legislative and regulatory efforts to reduce mercury contamination.
 - Anthropogenic mercury air emissions are decreasing from decreases in industrial uses and air pollution control devices.
 - Mercury has been found in POTW effluents despite extensive influent treatment.
 - Automobiles contribute approximately 0.75 to 1.5 short tons of mercury to nonhazardous waste landfills per year through auto shredder waste.
 - Of the 0.93 tons of mercury from Auto Shredder Waste, it is estimated that 0.4 short tons originated from automobiles.
 - Promote pollution prevention and recycling to provide additional safeguards from mercury environmental loading by regulating all mercury-containing waste as hazardous waste.

Endnotes

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- ⁶ Sznoppek and Goonan, 2000. p. 4.
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